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## Molecular genetic differentiation of native populations of Mediterranean blue mussels, *Mytilus galloprovincialis* Lamarck, 1819, and the relationship with environmental variables

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### Abstract

Blue mussels of the genus *Mytilus* are important ecosystem engineers in intertidal and subtidal communities. The distribution of *Mytilus* mussels is influenced by a number of benthic and pelagic environmental variables (e.g., substratum type and availability, water movement, phytoplankton production, physical disturbance) as well as interactions between these variables. Because of its broad tolerance of environmental variation the Mediterranean species, *Mytilus galloprovincialis*, has the greatest ability of all blue mussels to colonise new geographic regions. Understanding how population genetic variation is related to, or caused by, environmental variation is important but has long been a challenge. The present study examined the genetic differentiation of native populations of *M. galloprovincialis* throughout its entire geographic range in the Mediterranean Sea, the Black Sea and the Sea of Azov using 53 single nucleotide polymorphisms (SNP loci). Mussels, in total 1004 individuals collected from 36 locations, were genotyped and combined with existing SNP data for mussels from 11 reference sites. Pairwise comparisons of  $F_{ST}$  values, correspondence analysis (CA) and STRUCTURE analysis all revealed four groups of populations: the Atlantic Ocean; the western Mediterranean; the Aegean Sea; and the Azov, Black and Marmara Seas. One population – from Algeria (Oran West) – was intermediate between the two main groups of the Mediterranean Sea and Atlantic Ocean. Seascapes genetic analyses using GLM and DistLM analyses were employed to test site-specific genetic variation as a function of 13 environmental variables. The GLM identified five environmental variables that explained variation in site-specific  $F_{ST}$  values, whilst in the DistLM best-fit model only four were significant. These analyses suggest that a complex mix of environmental variables contribute to explaining genetic variation of *M. galloprovincialis* populations within the Mediterranean Sea, which most likely reflects the complex geological history of formation, isolation and reconnection among the regional sub-basins of the Sea.

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**Keywords:** Biogeography, single nucleotide polymorphism, seascape genetics, geological formation, isolation and reconnection

## Introduction

The Mediterranean Sea has a complex history of formation. Its origin dates to the Mesozoic and the Tethys Sea, when it acquired a shape similar to the present day in the Miocene (Picotti et al. 2014). At the end of the Miocene, the Mediterranean Sea was cut off from the Atlantic Ocean by a land barrier between Iberia and North Africa, which caused the water level to drop due to evaporation, which in turn gave rise to the Messinian crisis 5.75–5.32 million years before present (M ybp) (Rouchy & Caruso 2006; Janssen & Peijnenburg 2014). The connection to the Atlantic Ocean was re-established 5.3 M ybp when seawater filled the Mediterranean trough (Krijgsman et al. 1999, 2018). Subsequently, during the Pliocene, a monsoon climate period was followed by cyclical sea-level changes related to periods of glaciation in the Northern hemisphere (Barsotti & Meluzzi 1968; Picotti et al. 2014). Other major regions, sometimes considered to be a basin or sub-basin of the Mediterranean Sea, also have complex and sometimes completely independent geological histories. For example, the Black Sea, prior to the establishment of a contemporary connection to the Mediterranean Sea through the Bosphorus Strait, the Sea of Marmara (including the Dardanelle Strait) that was established 10.5 K ybp, was a freshwater lake receiving water from melting glaciers. Earlier, the Black Sea was only periodically influenced by Mediterranean Sea waters (Ryan et al. 2014). These complex geological histories and the present-day multi-basin structure of the Mediterranean Sea present an opportunity to better understand how periods of isolation and then reconnection, plus the different ages of the basins and their very different environmental conditions may influence genetic connectivity and population genetic diversity.

Given the complex history of formation of the Mediterranean Sea and its sub-basins it is not surprising that the Mediterranean Sea is a biodiversity hotspot (Coll et al. 2010; Danovaro et al. 2010; Pascual et al. 2017). Two main regions of biodiversity of the Mediterranean terrestrial fauna and flora have evolved, with different geological histories: the West Iberian and the North African separated by the Strait of Gibraltar to the west, and the Balkan and Anatolian to the east. These regions differ strongly in taxonomic composition and population divergencies of land organisms as a result of isolation in older (a few M ybp) and more recent glacial refugia (Sanmartín 2003; Froufe et al. 2016). In contrast to land and freshwater

organisms, marine species, including benthic sedentary organisms such as mussels, have the ability to disperse via their pelagic stages of development (larvae) and/or colonise new areas as adults on drifting items (e.g., kelp). Such dispersal potential, when realised, may mean that their populations are characterised by high connectivity and low spatial genetic differentiation (e.g., Lessios et al. 1998; Addison et al. 2008; Reisser et al. 2014). Consequently, such marine organisms may spread and reconnect isolated populations with the result that present-day marine populations in the Mediterranean basin may be only weakly genetically differentiated.

The native smooth-shelled blue mussel, *Mytilus galloprovincialis*, is a widely distributed species throughout the Mediterranean Sea. The species is an important ecosystem engineer in the intertidal and subtidal hard-bottom communities (Borthagaray & Carranza 2007; Arribas et al. 2014; Cinar et al. 2020), where the sessile adults are attached by byssus threads to hard substrata but the larvae may travel large distances in the water column. In addition, the mixing of Mediterranean populations may have been caused by a long history of shipping and trade activity associated with the sea (Blondel et al. 1999; Coll et al. 2010). For example, marine organisms including the mussel *Mytilus* have been used as food since the early Paleolithic (ca. 50 K ybp) in present-day southern Spain and Gibraltar (Cortés-Sánchez et al. 2019), the Neolithic in the Algarve, Portugal (Callapez & Pimentel 2018) and the Middle Bronze Age in present-day Italy (Zedda et al. 2003; Minniti 2005), as evidenced by preserved mussel shells in middens in caves and the vicinity of settlements. A simple form of mussel culture was initiated in Italy over 2 K ybp (Mattei & Pellizatto 1977; Smaal 2002). More recent intensive shipping activity, translocations of hatchery stocks and rafting on natural (e.g., kelp) and/or man-made (e.g., plastics) materials may have contributed to possible reduction of genetic differentiation among populations within basins of the Mediterranean Sea (Giantsis et al. 2014a).

The progenitor of modern smooth-shelled blue mussels gave rise to *Mytilus trossulus* in the Pacific Ocean, which subsequently invaded the North Atlantic Ocean after the opening of the Bering Strait 3.5 M ybp and gave rise to Atlantic Ocean *M. edulis* (Riginos & Cunningham 2005; Bach et al. 2019; reviewed by Gardner et al. 2021). This newly evolved taxon spread in the Northwest Atlantic Ocean to reach Europe. The oldest fossils

of *M. edulis* are found in deposits from the early Pliocene, its shells having been reported from middle and late Zanclean deposits in the vicinity of Lepe, on the Bay of Cadiz, Atlantic coast NW of Gibraltar (Belaústegui & Muñiz Guinea 2016). *M. edulis* colonised the Mediterranean Sea during periods of sea-level change before the Pleistocene (Vermeij 1992; Gardner et al. 2021). Glaciation in the Northern hemisphere took place 2.6 M ybp (Sosdian & Rosenthal 2009), likely resulting in a drop of sea level by 100 m. Divergence of *M. galloprovincialis* in the Mediterranean Sea from *M. edulis* in the North Atlantic Ocean probably occurred 2.5 M ybp and was followed by isolation lasting ~1.8 M y (Roux et al. 2014). Periods of glaciation forced the displacement of northern populations of boreal and colder-water species into southern Europe and resulted in the partial invasion of the Mediterranean Sea as a refugium, for example, by the bivalve *Arctica islandica* in the northern Adriatic Sea 1.8 M ybp (Crippa et al. 2016). During the Early–Middle Pleistocene Transition of 1.4 to 0.4 M ybp more regular cycles of glacial–interglacial periods became established and caused further changes to sea level. The global drop in sea level isolated the Mediterranean Sea biota, including its mussels, from the Northeast Atlantic Ocean biota. This scenario is supported by Riginos and Cunningham (2005) whose genetic analyses highlighted that isolation mechanisms (vicariance) were important for divergence of *M. galloprovincialis* from neighbouring *M. edulis*. It is possible that during such sea level changes the Atlantic Ocean mussel populations came into secondary contact with the Mediterranean Sea mussel populations many times in the period of (incomplete) separation that lasted ~1.7 M years (Barsotti & Meluzzi 1968). After colonisation by *Mytilus*, the influx of waters from the Atlantic Ocean to the present-day Mediterranean Sea was limited or cut off periodically due to tectonic changes, but mainly due to fluctuations in sea level to the west of the Strait of Gibraltar. In addition, strong evaporation of seawater caused the periodic lowering of the water level in the Mediterranean Sea and gave rise to the isolation of parts of it as separate basins during the Pleistocene (Bianchi et al. 2012). In the last 150,000 years, changes in sea level of ~130 m below to 6–15 m above the present sea level have been caused by glaciations and warming periods (Benjamin et al. 2017). This complex and dynamic geological history of the Mediterranean Sea has given rise to the allopatric divergence of *Mytilus* populations and the evolution of *M. galloprovincialis*. In addition, natural hybrid zones are created in the areas where modern

populations of *M. galloprovincialis* come into secondary contact with *M. edulis* (Bierne et al. 2003; Simon et al. 2020, 2021).

The ability of *Mytilus* mussels to spread on natural and artificial floating objects has contributed to their emergence in recent years in regions as far apart as Svalbard in the Arctic – *M. edulis* (Berge et al. 2005; Kotwicki et al. 2021) and the South Shetland Islands off Antarctica – *M. platensis* (Cárdenas et al. 2020). However, of all of the members of the genus it is *M. galloprovincialis* that shows the greatest ability to colonise new geographic regions, whether via human-mediated transfer (e.g., deliberately for aquaculture or accidentally via hull fouling or ballast water) or naturally (e.g., via kelp rafting) (Gardner et al. 2021). In recent years, it has been recorded on the Atlantic coasts of South America, where in Argentina it hybridises with the native taxon, *M. platensis* (Zbawicka et al. 2018), and in Brazil where it inhabits farmed cultures of the native species, *Perna perna* (Birckolz et al. 2020; Lins et al. 2021). *M. galloprovincialis* has also been introduced to South Africa, the Pacific coast of North America, the Sea of Japan, China, Korea, Australia, New Zealand and Chile (Wilkins et al. 1983; McDonald & Koehn 1988; McDonald et al. 1991; Gardner et al. 2016; Han et al. 2017; Larraín et al. 2018; Zbawicka et al. 2019, 2022; Popovic et al. 2020). It is considered to be one of the most successful invasive species of the global coastal marine biota, often displacing native species (Geller 1999; Lowe et al. 2000; Gardner et al. 2021).

The distribution of *Mytilus* species is controlled by a number of environmental processes involving both benthic and pelagic habitats (e.g., substratum type and availability, water movement, phytoplankton production, physical disturbance, seawater biogeochemistry) as well as interactions between these processes (Sandman et al. 2013; Kotta et al. 2015). For example, water movement can indirectly affect *Mytilus* spp. by modifying sedimentation rates or may directly affect sessile mussels by physically disturbing or detaching them (Westerbom et al. 2008). Furthermore, the sedentary benthic suspension feeding lifestyle of adult mussels is a life-history characteristic that requires water-borne food delivery, but the quantity and quality of such food delivery is highly variable in time and space (Dahlhoff & Menge 1996; Gardner 2000, 2013; Saurel et al. 2007). An ability to withstand environmental variability is often key to a species' range expansions, whether they be natural or human-mediated. Environmental tolerance can be related to or controlled by such mechanisms of adaptation as differences in gene family expression, gene splicing,

methylation, past gene duplications and other genomic mechanisms also related to pangenome functions (Pujolar et al. 2014; Malachowicz et al. 2015; Kijewska et al. 2016, 2018; Bitter et al. 2019; Malachowicz & Wenne 2019; Clark et al. 2021; Corrochano-Fraile et al. 2022; Liu et al. 2022). For example, there is now a growing body of evidence to indicate that *M. galloprovincialis* is the best adapted smooth-shelled mussel to tolerate thermal variation, an adaptation that may be both facilitating its spread into non-native areas and its ability to outcompete congeners (e.g., Braby & Somero 2006; Jones et al. 2010; Tomanek & Zuzow 2010). When seeking to understand associations between environmental and population genetic variation it is important to incorporate ecologically meaningful drivers into the models (e.g., direct and indirect environmental gradients, resource gradients). However, trying to identify environmental drivers that are relevant at different temporal and spatial scales is often difficult because the drivers may not act in isolation but may instead act synergistically, although not always to the same extent in time (e.g., different seasonal interactions) or in space (e.g., coastal headlands versus bays). Thus, quantifying interactive effects (i.e., how different environmental gradients separately or interactively modulate the resource–allele relationship) is still a major challenge (Palumbi et al. 2019; Hu et al. 2020; Zeng et al. 2020).

To date, genetic studies of native *M. galloprovincialis* populations in the Mediterranean Sea are local in nature, having covered only certain regions (e.g., Ahmad & Beardmore 1976; Edwards & Skibinski 1987; Karakousis & Skibinski 1992; Quesada et al. 1995a, 1995b; Comesáñ et al. 1998; Daguin & Borsa 1999; Giantsis et al. 2012, 2014a, 2014b; Bierne et al. 2003; Sammer et al. 2010; Vera et al. 2010; Lourenço et al. 2015; Fraïsse et al. 2016; Paterno et al. 2019). No genetic research has yet been undertaken at the pan-Mediterranean Sea scale. The aim of the present study was to test for genetic differentiation of *M. galloprovincialis* populations throughout its entire native range. We also tested if and how much environmental variables explained the population genetic diversity of *M. galloprovincialis*. By including key environmental variables in our testing we seek to explore links between patterns of environmental and genetic variability of the mussel populations. A strong match between genetic and environmental variation and the presence of clear genetic differentiation according to present day environmental gradients may indicate the importance of local adaptation processes in Mediterranean Sea *M. galloprovincialis*

or barriers to gene flow. Alternatively, a relationship between environmental variation and subtle regional genetic differentiation may reflect the complex geological history of the region and the evolutionary history of the mussel populations.

## Materials and methods

### *The environment*

The Mediterranean Sea is a land-locked, semi-enclosed marginal sea of the Atlantic Ocean. The Mediterranean Sea is a very complex marine environment and is oceanographically diverse with a number of distinct sub-seas. Sea surface temperature exhibits strong seasonality and pronounced spatial gradients from west to east and north to south. The Mediterranean Sea is microtidal with a typical tidal range of less than 50 cm. The salinity of the Mediterranean Sea is high throughout the basin from 35 PSU on the West to 39 PSU on the East, but is brackish in the Black Sea and Sea of Azov. The Mediterranean Sea has very low nutrient concentrations, especially in its eastern parts, which is reflected in the patterns of primary productivity. Locally, riverine nutrient input can be significant; however, most river systems discharging into the Mediterranean Sea are small by global standards (Uitz et al. 2012; Goffredo & Dubinsky 2014).

### *Sampling and isolation of genomic DNA from mussel samples*

*Mytilus* spp. samples, consisting in total of 1004 individuals of mixed ages (not quantified) and sizes (5 to 50 mm shell length) were collected from 36 sites (nominally “populations”) situated along the coasts of the Atlantic Ocean, the Mediterranean Sea and the Black Sea between 2004 and 2016 (Table I; Figure 1). DNA was isolated from mantle tissue that had been stored in 96% ethanol or at -70°C, using a modified CTAB method (Hoarau et al. 2002). Eleven reference site samples of *M. galloprovincialis* from the Atlantic Ocean coast of Spain and Portugal, the Mediterranean Sea and the Sea of Azov, *M. edulis* from France and *M. trossulus* from Canada, consisting of 316 specimens, were also included (Table I). Fifty three single nucleotide polymorphisms (SNPs) differentiating *Mytilus* taxa, and with the ability to identify hybridisation, were employed to assay genetic variation (Zbawicka et al. 2012, 2018; Gardner et al. 2016). Samples were genotyped using the Sequenom MassARRAY iPLEX genotyping platform (Gabriel et al. 2009).

Table I. Sample code of *Mytilus*, numbers of specimens examined (N), location and country, geographic coordinates and year of sampling of all mussel populations.

No.	Sample	N	Location	Country	Coordinates	Year
1	AGA	31	Agadir, Atlantic	Morocco	30°18'3.26"N	9°48'56.60"W
2	ALC	29	Alcanar, Balearic S.	Spain	40°34'8.51"N	0°32'30.07"E
3	BAN	27	Banyuls, Gulf of Lion	France	42°27'51.89"N	3°10'30.49"E
4	BGT	30	Bizerta Bay, Gulf of Tunis	Tunisia	37°16'36.70"N	9°53'58.20"E
5	BLT	30	Bizerta Lagoon	Tunisia	37°10'30.89"N	9°49'41.04"E
6	BLS	30	Crimea	Ukraine	44°29'0.82"N	34°21'8.92"E
7	BRI	30	Brindisi, South Adriatic S.	Italy	40°38'30.82"N	17°57'56.69"E
8	CAT	30	Catholica, South Adriatic S.	Italy	43°59'36.15"N	12°44'49.03"E
9	CHW	35	Istria, North Adriatic S.	Croatia	45°751.44"N	13°39'39.81"E
10	CIRP	28	Mare Piccolo, Taranto, North Ionian S.	Italy	40°30'2.00"N	17°18'1.00"E
11	DUB	34	Dubrovnik, East-Central Adriatic	Croatia	42°32'49.39"N	18°10'21.69"E
12	EGE	29	Porto Covo, Aegean S.	Greece	39°57'49.00"N	23°55'19.00"E
13	GAE	30	Gulf of Gaeta, Thyrrenian S.	Italy	41°13'39.84"N	13°34'6.72"E
14	GAR	30	Gargano (buoy), South Adriatic	Italy	41°21'27.56"N	16°17'15.81"E
15	GRK	30	Kassiopi, Corfu, North Ionian S.	Greece	39°47'31.32"N	19°52'12.66"E
16	GRM	30	Mazoma Lagoon, North Ionian S.	Greece	39° 0'52.83"N	20°44'44.93"E
17	HER	30	Heraklion, Crete, South Aegean S.	Greece	35°20'40.96"N	25° 8'56.50"E
18	IMC	20	Torre Grande Marina, Oristano, Sardinia	Italy	39°54'4.42"N	8°29'4.18"E
19	LAMW	30	L'Ampolla wild, Balearic S.	Spain	40°46'13.28"N	0°45'28.88"E
20	LAS	16	La Spezia, Ligurian S.	Italy	44° 5'29.43"N	9°49'45.10"E
21	MAK	29	Makarska, East-Central Adriatic S.	Croatia	43°16'34.99"N	17° 0'51.84"E
22	MOM	30	Canakkale, Dardanelle Strait	Turkey	40° 9'21.17"N	26°24'28.28"E
23	NEA	30	Gulf of Naples, Thyrrenian S.	Italy	40°46'44.64"N	14° 5'28.20"E
24	ORAE	30	Oran East, Alboran S.	Algeria	35°42'33.74"N	0°39'14.64"W
25	ORAW	29	Oran West, Alboran S.	Algeria	35°10'44.16"N	1°38'57.67"W
26	ORB	25	Orbetello, Thyrrenian S.	Italy	42°26'21.00"N	11°13'14.00"E
27	PIST	27	Pisticci, Gulf of Taranto, North Ionian S.	Italy	40°20'12.00"N	16°2'31.00"E
28	PSL	12	Port-Saint-Louis-du-Rhône	France	43°22'55.40"N	4°52'46.86"E
29	SAR	30	Saronikos Gulf, Aegean S.	Greece	37°35'39.64"N	23°16'58.52"E
30	SBRB	29	Sunny Beach, South-East Black S.	Bulgaria	42°41'58.74"N	27°43'3.38"E
31	SET	23	Sete, Gulf of Lion	France	43°23'27.30"N	3°41'48.11"E
32	SGL	23	Santa Giusta Lagoon (Stagno), Sardinia	Italy	39°43'39.27"N	8°29'49.91"E
33	SIS	31	Sisitana, North Adriatic S.	Italy	45°46'19.19"N	13°35'11.26"E
34	TES	30	Thessaloniki, Aegean S.	Greece	40°38'1.23"N	22°52'31.97"E
35	TURK	18	Izmir, Aegean S.	Turkey	38° 4'26.33"N	27° 1'19.61"E
36	VAL	29	Valencia, Balearic S.	Spain	39°26'57.08"N	0°18'39.11"W
37	AZO <sup>a</sup>	30	Sea of Azov	Ukraine	45°43'51.71"N	35° 5'0.26"E
38	LID <sup>a</sup>	32	Iido	Italy	45°30'25.42"N	12°38'30.04"E
39	ORI <sup>a</sup>	29	Oristano	Italy	39°47'59.88"N	8°31'9.72"E
40	CAM <sup>b</sup>	29	Camarinal	Spain	36°4'48.01"N	5°47'58.00"W
41	CAS <sup>b</sup>	30	Cascais	Portugal	38°34'14.89"N	9°19'8.95"W
42	VIG <sup>b</sup>	30	Vigo	Spain	42°13'54.12"N	8°45'7.22"W
43	BID <sup>b</sup>	29	Bidasoa	Spain	43°21'38.71"N	1°51'11.15"W

(Continued)

No.	Sample	N	Location	Country	Coordinates	Year
44	<b>CHA<sup>c</sup></b>	20	Charente	France	45°56'60.00"N 1°15'20"W	2015
45	<b>LOI<sup>c</sup></b>	29	Loire	France	47°14'43.83"N 2°13'48.88"W	2004
46	<b>MSMA<sup>c</sup></b>	30	Mont Saint Michel Bay	France	48°39'0.06"N 1°31'40.26"W	2015
47	<b>KKAT<sup>d</sup></b>	28	Halifax	Canada	44°30'33.79"N 63°29'24.91"W	1996

Reference samples are marked in bold (<sup>a</sup>*M. galloprovincialis* Mediterranean; sample 37 - Wenne et al. 2020, 38 - 2016, 39 - Zbawicka et al. 2018; <sup>b</sup>*M. galloprovincialis* Atlantic: 40 -, 2018, 41 and 42 -; Wenne et al. 2020, 43 -, 2016; <sup>c</sup>*M. edulis*: 44 - Simon et al. 2021, 45 - Wenne et al. 2016, 46 - unpublished data; <sup>d</sup>*M. trossulus*: 47 - Zbawicka et al. 2018).

### Data analysis

**Genetic diversity.** Population genetic variation was analysed in Arlequin v. 3.5.1.2 (Excoffier & Lischer 2010) to estimate: locus-specific allele frequencies, proportion of polymorphic SNPs ( $P_O$ ), minor allele frequencies (MAF for bi-allelic), observed ( $H_O$ ) and expected ( $H_E$ ) heterozygosities, genetic diversity and differentiation (pairwise  $F_{ST}$ ), inbreeding coefficient ( $F_{IS}$ ) and departures from Hardy-Weinberg equilibrium (HWE). An analysis of molecular variance (AMOVA) with 1000 permutations to estimate variance components among *Mytilus* populations was also conducted in Arlequin. The most informative loci were identified by calculating  $F_{ST}$  outlier loci using the approach of Excoffier et al. (2009) implemented in Arlequin. The upper limits of the 95% confidence intervals were identified with 20,000 iterations and characterised by estimating the 0.05 and 0.95 and 0.99 quantiles of the distribution. The false discovery rate (FDR-BY) was applied to correct significance (P) values after multiple testing (Benjamini & Yekutieli 2001; Narum 2006).

**Population genetic differentiation and structure.**  $F_{ST}$  measures in the Newick format, obtained in POPTREEW (Takezaki et al. 2014), were used to construct a neighbour-joining (NJ) tree illustrating the differentiation among populations (all *M. galloprovincialis* and *M. edulis*, but not the reference *M. trossulus*) and visualised in MEGA version 6 (Tamura et al. 2013). Two methods were used for the population structure analyses. First, correspondence analysis (CA; Benzécri 1992), implemented in GENETIX (Belkhir et al. 2003), was used to visualise genetic structure among populations. The results are presented as a scatter plot, with the axes representing the contribution of inertia of the data matrix in a way that can be considered analogous to the total variance in allelic frequencies (Benzécri 1992). Second, clustering and assignment testing were performed using the Bayesian-based method implemented in STRUCTURE v. 2.3.4. STRUCTURE was employed using the model assuming admixture, ignoring population affiliation and allowing for the correlation of allelic frequencies between clusters. This admixture model allows for individual structure with mixed ancestry, meaning that fractions of the genome could have come from different ancestors (Pritchard et al. 2000; Falush et al. 2007). The most appropriate number of genetic clusters ( $K$ ) was determined by a diagram-based comparison of log-likelihoods for values of  $K$ . At least five runs were used to calculate each  $K$  value, following the method described by Evanno et al.

Table I. (Continued).



Figure 1. Locations of 36 sampling and 10 reference sites in the Mediterranean Sea and Atlantic Ocean (KKAT from Halifax, Canada, Northwestern Atlantic Ocean is not shown). Sample site names and geographical coordinates are detailed in [Table I](#).

(2005). Threshold  $q$ -values of 0.2 were used as a criterion to separate hybrid and pure mussels (Vähä & Primmer 2006). Individuals were considered residents if  $q > 0.8$  for the area where they were sampled. Individuals with  $q$ -values from 0.2 to 0.8 were considered to be potentially admixed, as they could not be readily assigned as residents or migrants (Lecis et al. 2006). Individuals with an assignment probability of  $q > 0.8$  were defined as belonging to the wild population (cf. Jonker et al. 2013), whilst those with  $q \leq 0.8$  were labelled as admixed. A Monte Carlo Markov Chain was run for 100,000 iterations following a burn-in period of 50,000 iterations.

**Linking genetic and environmental variation.** Regional-scale proxies for weather and local environmental conditions were obtained from online databases for 13 environmental variables (Table S1). As a proxy for the global climate and weather (air and water temperature, cloud, wind speed, waves) we used ERA5 reanalyses (for all methodological details of environmental variables and modelling see references in Table S1).

Reanalysis combines model data with observations from across the world into a globally complete and consistent dataset. ERA5 provides hourly estimates for a large number of atmospheric, ocean-wave and land-surface variables. To characterise biogeochemical patterns (concentrations of chlorophyll *a*, nitrate and phosphate, as well as marine primary production) we used the CMEMS global biogeochemical multi-year hindcast product GLOBAL REANALYSIS BIO 001 029. CMEMS provides daily estimates for all biogeochemical variables. Model inputs for the physical conditions (salinity, mixed layer thickness) were obtained from the CMEMS global ocean eddy-resolving reanalysis product GLOBAL REANALYSIS PHY-001-030. Data for the global distribution of photosynthetically available radiation at the sea surface and on the seafloor were obtained from Gattuso et al. (2020).

Relationships between environmental variables and genetic variation of *M. galloprovincialis* populations from the Mediterranean Sea (i.e., excluding populations CHA, LOI, MSMA and KKAT) were tested using two complementary seascape genetics approaches. First, following Wei et al. (2013), we

used population-specific  $F_{ST}$  values (the response variable and where negative  $F_{ST}$  values were set to zero) and the 13 environmental variables in a GLM (the GLZ routine in Statistica v. 12). We employed a normal distribution and a log link function. To minimise Type I error that may be associated with stepwise (forward or backward) model building we employed the “all effects” model. Second, following Silva and Gardner (2016), we used population-specific allelic frequencies (the response variables) and the 13 environmental independent variables in a distance-based linear model (DistLM in the PRIMER + PERMANOVA v. 6 software package - Anderson et al. 2008). This test is a permutational equivalent to partial redundancy analysis (Legendre & Anderson 1999). DistLM was used to perform an ordination of fitted values from a given model and is constrained to find linear combinations of predictor variables (environmental data) that explain the greatest variation in the data cloud (population-specific allele frequencies). Permutation of residuals was carried out under a reduced (or partial) model and because this is a permutational test, there are no assumptions about data normality (Anderson et al. 2008). Marginal tests (one independent variable at a time) and sequential tests (all independent variables entered into the model based on their relative importance (most significant first) in the marginal tests) were employed to identify the environmental variables that explained the greatest variation in the genetic dataset. Model fit was tested using adjusted  $R^2$  (i.e., adjusted for the number of terms in the model), the AIC value and the BIC value. Note that these two seascape genetic analyses use the same environmental dataset but test for population-specific variation in different dependent variables: for the GLM this is population-specific  $F_{ST}$  values (a summary metric of population differentiation) whereas for the DistLM this is locus-specific allele frequencies (i.e., the raw genetic data).

## Results

### SNP validation and Hardy-Weinberg equilibrium

Fifty-three SNPs were genotyped (Table II). ORF identification was not possible for only five SNPs. Forty-four SNPs (83.01%) were located in coding regions, of which only three (5.66%) were not synonymous. Four SNPs (7.54%) were located in non-coding regions. The MAF ranged from 0.000 (7 different loci) to 0.404 (1 locus) (Table II). The vast majority of loci were in Hardy-Weinberg equilibrium (HWE) in all populations.

No SNPs with departures from HWE were observed in 10 populations (AGA, BLS, CHW, CIRP, MOM, ORAW, SBRB, TURK and reference CAS and LOI). Only one population (VAL) had three SNPs that were not in HWE ( $P < 0.01$ ). In the remaining samples the fraction of SNPs showing significant departures from HWE was one or two for 20 and 14 populations, respectively (Table III).

### Detection of outlier loci and highly informative SNPs

Eighteen of the 53 SNPs differentiated *M. trossulus* (Halifax, Nova Scotia, Canada) from other taxa (BM33B, BM44B, BM30C, BM201B, BM12A, BM103B, BM10B, BM30A, BM11A, BM2G, BM101A, BM116A, BM64A, BM202B, BM202A, BM92B, BM203D, BM62A). When the *M. trossulus* reference sample was excluded from the analysis seven outlier SNPs differentiating *M. edulis* from *M. galloprovincialis* (BM101A, BM12C, BM17B, BM21C, BM201C, BM57A, BM67C) were identified (Fig. S1). After excluding reference samples of *M. edulis* and the Atlantic Ocean lineage of *M. galloprovincialis* an analysis was performed of the samples from the Mediterranean Sea and the Black Sea (including the Sea of Azov, the Marmara Sea and the Straits of Dardenelle). Two outlier SNPs differentiating Mediterranean Sea from Black Sea/Sea of Azov *M. galloprovincialis* populations (BM105A, BM106B) were identified (Fig. S2).

### Genetic diversity

The percentage of polymorphic loci ( $P_o$ ) in all *M. galloprovincialis* populations ranged from 45% in the Morocco (AGA) and Turkey (TURK) samples to 56.60% in the Croatia (DUB) and reference IMC (Italy) samples (Table III). Observed heterozygosity ( $H_o$ ) for 53 loci across 47 populations was lower than expected ( $H_e$ ), except for two samples: ORAE and BLS (Table III). The highest  $H_o$  values ranged from 0.310 to 0.327 and were recorded at TURK, ORAW, AGA and ORAE. The highest values of the expected heterozygosity ( $H_e$ ) were observed at SET, TURK and AGA (0.327, 0.328 and 0.341, respectively), whilst the lowest values were observed at PIST and TES (0.278 and 0.287, respectively). The mean within-population fixation index  $F_{IS}$  (averaged over all polymorphic loci in 47 populations) was 0.069. For seven samples (GRM, SIS, CIRP, GAR, BAN, VAL and reference *M. trossulus* KKAT) significant  $F_{IS}$  values indicated an excess of homozygotes (Table III),

Table II. Descriptive properties of 53 SNP loci among 47 populations of *Mytilus* spp., genome location, substitution type, allelic location, substitution type, allelic variant,  $F_{ST}$  P-value associated with test for outlier status, minor allele frequency (MAF), GenBank accession number and supporting reference.

Locus name	Location	Region	Substitution	Allele	$F_{ST}$	$H_O$	$F_{IS}$	No. of sampling sites where the locus was polymorphic	MAF	GenBank accession number	Reference
1 BM101A	Ribosomal protein L7a	coding	synon	A/T	<b>0.957</b>	0.159	0.210	5	0.003	KT713378	Wenne et al. 2016
2 BM102A	Proteasome subunit beta type-6-like	coding	synon	C/T	<b>0.049</b>	0.295	-0.043	46	0.177	KT713379	Wenne et al. 2016
3 BM103B	Proteasome subunit beta type-5-like	coding	nonsyn	A/G	<b>0.736</b>	0.032	0.054	1	0.006	KT713380	Wenne et al. 2016
4 BM105A	UnKnown	NA	NA	A/G	<b>0.110</b>	0.492	0.326	45	0.384	KT713381	Wenne et al. 2016
5 BM106B	UnKnown	NA	NA	A/G	<b>0.088</b>	0.217	0.996	33	0.137	KT713382	Wenne et al. 2016
6 BM10B	Ribosomal protein S20	coding	synon	A/C	<b>0.805</b>	0.035	0.027	1	0.004	KJ871040	Zbawicka et al. 2014
7 BM116A	ATP synthase beta subunit	coding	synon	A/C	<b>0.965</b>	0.047	0.000	2	0.001	KT713387	Wenne et al. 2020
8 BM118A	UnKnown	NA	NA	A/G	<b>0.028</b>	0.274	0.017	46	0.168	KT713388	Wenne et al. 2016
9 BM11A	Ribosomal protein L22	coding	synon	A/G	<b>0.931</b>	0.045	-0.008	3	0.001	KJ871041	Zbawicka et al. 2014
10 BM121A	Ribosomal protein L30	coding	synon	A/T	0.000	0.001	0.000	1	0.000	KT713389	Wenne et al. 2020
11 BM12A	Ribosomal protein L23a	coding	synon	C/T	<b>0.750</b>	0.044	-0.060	7	0.006	KJ871042	Zbawicka et al. 2014
12 BM12C	Ribosomal protein L23a	coding	synon	C/T	<b>0.235</b>	0.482	0.034	47	0.255	KJ871042	Wenne et al. 2016
13 BM147A	UnKnown	NA	NA	C/T	<b>0.114</b>	0.312	0.048	45	0.175	KT713383	Wenne et al. 2016
14 BM16B	ribosomal protein L21	coding	synon	C/T	<b>0.021</b>	0.396	0.044	47	0.268	KJ871044	Wenne et al. 2016
15 BM17B	Ribosomal protein L7a	coding	synon	A/G	0.516	0.090	0.089	15	0.030	KJ871045	Zbawicka et al. 2014
16 BM201B	<i>H4 histone gene</i>	noncoding	NA	A/C	<b>0.667</b>	0.032	0.616	3	0.007	AY267750.1	Zbawicka et al. 2012
17 BM201C	<i>H4 histone gene</i>	coding	synon	G/T	<b>0.214</b>	0.417	0.001	45	0.216	AY267750.1	Zbawicka et al. 2012
18 BM202A	<i>H3 histone gene</i>	noncoding	NA	A/C	1.000	1.000	0.000	0	0.000	AY267749.1	Zbawicka et al. 2012
19 BM202B	<i>H3 histone gene</i>	coding	synon	A/T	1.000	1.000	0.000	0	0.000	AY267749.1	Zbawicka et al. 2012
20 BM203B	<i>H2B histone gene</i>	noncoding	NA	C/T	<b>0.193</b>	0.442	0.054	45	0.246	AY267742.1	Zbawicka et al. 2012
21 BM203D	<i>H2A histone gene</i>	noncoding	NA	A/T	1.000	1.000	0.000	0	0.000	AY267757.1	Zbawicka et al. 2012
22 BM204A	<i>p53</i>	coding	synon	C/T	<b>0.104</b>	0.498	-0.006	46	0.364	DQ865151	Zbawicka et al. 2012
23 BM21B	qm-like protein	coding	nonsyn	C/G	<b>0.982</b>	0.043	0.000	1	0.000	KJ871047	Zbawicka et al. 2014

Table II. (Continued).

Locus name	Location	Region	Substitution	Allele	$F_{ST}$	$H_O$	$F_{IS}$	was polymorphic	MAF	No. of sampling sites where the locus	GenBank accession number	Reference
24 BM21C	qm-like protein	coding	synon	A/C/ T	<b>0.237</b>	0.378	-0.035	47	NA	KJ871047	Zbawicka et al. 2014	
25 BM26B	UnKnown13	NA	NA	A/T	<b>0.190</b>	0.215	0.000	41	0.105	KJ871050	Zbawicka et al. 2014	
26 BM2G	UnKnown05	coding	synon	G/T	<b>0.944</b>	0.041	-0.041	1	0.001	KJ871032	Zbawicka et al. 2014	
27 BM30A	Ribosomal protein L17	coding	synon	A/G	<b>0.855</b>	0.037	-0.149	1	0.003	KJ871052	Zbawicka et al. 2014	
28 BM30C	Ribosomal protein L17	coding	synon	A/T	<b>0.491</b>	0.023	0.082	2	0.011	KJ871052	Zbawicka et al. 2014	
29 BM32A	ubiquinol-cytochrome c reductase subunit 6	coding	synon	A/G	<b>0.032</b>	0.493	0.406	47	0.402	KT713371	Wenne et al. 2016	
30 BM33B	Cytochrome c oxidase subunit IV	coding	synon	A/T	<b>0.453</b>	0.020	-0.327	1	0.010	KJ871054	Zbawicka et al. 2014	
31 BM35C	Ribosomal protein L7	coding	synon	A/T	<b>0.102</b>	0.368	0.036	47	0.215	KJ871055	Wenne et al. 2016	
32 BM35D	Ribosomal protein L7	coding	synon	A/G	<b>0.084</b>	0.292	0.128	45	0.168	KJ871055	Zbawicka et al. 2014	
33 BM36F	ribosomal protein S3a	coding	synon	A/C	<b>0.215</b>	0.170	-0.003	46	0.071	KT713373	Wenne et al. 2016	
34 BM38B	ribosomal protein S8e	coding	synon	A/G	<b>0.045</b>	0.302	0.054	47	0.172	KT713368	Wenne et al. 2016	
35 BM44B	ubiquitin/ribosomal protein S27a	coding	synon	A/G	0.490	0.025	0.430	3	0.011	KJ871057	Zbawicka et al. 2014	
36 BM50B	CoA-binding protein	coding	synon	A/G	0.056	0.003	-0.059	1	0.002	KJ871059	Zbawicka et al. 2014	
37 BM54A	ETC C1 NDUFA4	coding	synon	A/G	0.672	0.063	0.065	20	0.010	KJ871060	Zbawicka et al. 2014	
38 BM57A	NADH-ubiquinone_oxidoreductase	coding	nonsyn	C/T	<b>0.199</b>	0.442	0.028	45	0.242	KT713374	Wenne et al. 2016	
39 BM57D	NADH-ubiquinone oxidoreductase	coding	synon	A/C	<b>0.155</b>	0.470	-0.003	46	0.290	KT713374	Wenne et al. 2016	
40 BM5B	Ribosomal protein S6e	coding	synon	A/G	<b>0.058</b>	0.207	0.023	41	0.114	KJ871035	Wenne et al. 2016	
41 BM5D	Ribosomal protein S6e	coding	synon	C/T	<b>0.035</b>	0.355	-0.004	47	0.227	KJ871035	Zbawicka et al. 2014	
42 BM60A	UnKnown08	coding	synon	A/G	0.192	0.013	0.051	4	0.006	KJ871063	Zbawicka et al. 2014	
43 BM61A	Ribosomal_L1	coding	synon	C/T	<b>0.134</b>	0.211	0.037	45	0.106	KT713375	Wenne et al. 2016	

(Continued)

Table II. (Continued).

Locus name	Location	Region	Substitution	Allele	$F_{ST}$	$H_O$	$F_{IS}$	No. of sampling sites where the locus was polymorphic	MAF	GenBank accession number	Reference
44 BM62A	Ribosomal L13e	coding	synon	A/G	1.000	1.000	0.000	0	0.000	KJ871064	Zbawicka et al. 2014
45 BM64A	Ribosomal protein L35	coding	synon	C/T	<b>0.982</b>	0.044	0.001	1	0.000	KJ871065	Zbawicka et al. 2014
46 BM67C	Ribosomal protein S6e	coding	synon	A/T	<b>0.252</b>	0.312	0.226	47	0.154	KJ871066	Wenne et al. 2016
47 BM6C	EFG N	coding	synon	C/T	0.273	0.029	0.124	8	0.015	KJ871036	Zbawicka et al. 2014
48 BM75C	UBA UBA-TS-N domain	coding	synon	C/G	0.047	0.003	-0.049	1	0.001	KT713370	Wenne et al. 2016
49 BM78B	UnKnown12	coding	synon	A/G	<b>0.125</b>	0.437	0.062	47	0.262	KJ871069	Zbawicka et al. 2014
50 BM8E	Ribosomal protein L3	coding	synon	A/G	0.546	0.079	0.067	29	0.019	KJ871038	Zbawicka et al. 2014
51 BM92B	UnKnown06	coding	synon	A/T	1.000	1.000	0.000	0	0.000	KJ871074	Zbawicka et al. 2014
52 BM9B	Ribosomal protein S2	coding	synon	A/G	<b>0.075</b>	0.297	0.037	47	0.175	KJ871039	Zbawicka et al. 2014
53 BM9C	Ribosomal protein S2	coding	synon	A/C/ T	<b>0.132</b>	0.401	0.031	47	NA	KJ871039	Zbawicka et al. 2014

$F_{ST}$  - fixation index, values with  $P < 0.05$  after Benjamini-Yekutieli corrections are marked in bold;  $H_O$  - observed heterozygosity;  $F_{IS}$  - inbreeding coefficient; No. - number of sampling sites where the locus was polymorphic; MAF - minor allele frequency.

which may indicate a relationship between individuals within a population, resulting from the collection of related individuals from a small area.

Estimates of average pairwise differences within-population diversity among the *M. galloprovincialis* populations revealed that the most diverse population was BLT from Tunisia and the least diverse was CHW from Croatia. In the reference samples the greatest diversity was observed in the Atlantic *M. galloprovincialis* BID population from Spain and the lowest in the *M. edulis* MSMA population from France. The same samples exhibited the greatest and the least gene diversity ([Table III](#)).

Allelic frequencies were calculated for 53 SNPs and minor allele frequency (MAF) was determined for 51 bi-allelic SNPs ([Table S2](#)). Across all 47 populations the MAF values ranged from 0.075 to 0.124. MAF values in the *M. galloprovincialis* populations ranged from 0.087 to 0.124, and were generally low, but higher than for *M. edulis* reference populations: CHA (0.075), LOI (0.083) and MSMA (0.083). MAF values for the *M. galloprovincialis* samples from the Black Sea (0.087) and the Aegean Sea (0.094) were low in comparison to other *M. galloprovincialis* samples. The MAF values for Atlantic *M. galloprovincialis* were higher (0.110, 0.119, 0.122 and 0.124 for CAS, CAM, BID and VIG, respectively - [Table III](#)) than for the Mediterranean Sea populations of *M. galloprovincialis*.

#### *Genetic differentiation among populations*

Pairwise comparisons of  $F_{ST}$  values among all 47 populations for the 53 SNPs indicated that many pairs of populations were significantly different from one another ( $P < 0.05$  after Benjamini-Yekutieli correction) ([Table S3](#)). As expected, statistically significant genetic differentiation was observed for most of the pairwise comparisons involving reference *M. trossulus* (KKAT), reference Atlantic Ocean *M. galloprovincialis* (BID, VIG, CAS and CAM) and reference *M. edulis* (MSMA, LOI and CHA). The *M. galloprovincialis* populations AGA, MOM, ORAW and BLS exhibited the greatest numbers of statistically significant values of  $F_{ST}$ . Low but statistically significant levels of differentiation were observed between pairs of populations from the Aegean Sea and the Black Sea. The highest  $F_{ST}$  values were observed in the pairwise comparisons of SRRB (the Black Sea) with SAR (0.077) and TURK (0.073), both in the Aegean Sea, as well as between AZO (the Sea of Azov) and TURK (0.075). The pairwise  $F_{ST}$  values for the Aegean Sea and the Black Sea populations were also

statistically significant when compared to other Mediterranean Sea populations: the highest value was observed for the TURK and SGL pair (Santa Giusta Lagoon, Italy) and was 0.061. Most other pairwise  $F_{ST}$  values were  $\leq 0.03$  and were not significantly different from zero. There was no statistically significant differentiation for MOM (Dardanelle Strait) and populations from the Black Sea (SRRB, BLS, AZO), but differentiation between MOM and most populations from the Aegean was observed: MOM-HER (0.047), MOM-SAR (0.048), MOM-TES (0.040) and MOM-TURK (0.047). For the pairwise comparison of MOM-EGE the  $F_{ST}$  value was not-significantly different from zero, indicating the absence of genetic differentiation. No significant genetic differentiation was observed for the CHW (northern Adriatic Sea) – BLS (Black Sea) pair.

AMOVA was performed comparing groups of samples for five different scenarios where populations were defined *a priori* (details in [Table IV](#)). The estimated values of the  $F$ -statistic were significant for all five scenarios, and the greatest variance was exhibited within individuals for all scenarios. Considering samples from the Mediterranean Sea as one group the percent of variation was the greatest (~21%) among populations in comparison not with other scenarios (range ~1.2 to 3.2%). The levels of percent variance were quite similar when the samples from the Mediterranean Sea basin were divided into 2 or 4 groups. AMOVA analysis with the division into two groups (the Atlantic Ocean and the Mediterranean Sea with the Black Sea and the Sea of Azov) showed a significant increase of the variance component value among populations within groups. The inclusion of the ORAW population from the Alboran Sea in both the Atlantic Ocean and Mediterranean Sea groups did not affect the differences in proportions of genetic variability.

Neighbour-joining (NJ) tree analysis with *M. edulis* as the outgroup revealed relationships among *M. galloprovincialis* populations ([Figure 2](#)). A separate group with 97% bootstrap support consisted of reference samples of Atlantic Ocean *M. galloprovincialis* (BID, CAM, CAS and VIG), with the Moroccan Atlantic Ocean sample (AGA) plus the ORAW (inside the Mediterranean Sea) sample. This group was clearly separated from all other *M. galloprovincialis* populations. Within the *M. galloprovincialis* from the Mediterranean Sea, two main branches of the tree were distinguished: the first encompassed populations from the Sea of Azov, the Black Sea, and the Marmara Sea (MOM, SRRB, BLS, AZO), and the second from around the Aegean Sea (EGE, HER, SAR, TURK, TES and BRI). Between these two clusters several

Table III. Genetic diversity indices of the 47 *Mytilus* populations.

Sample	$P_O$	$H_O$	$H_E$	MAF	Av. gene diversity over loci	Average no. pairwise differences within population	No. loci with departure from HWE; P < 0.01	$F_{IS}$
AGA	45.28	0.318	0.341	0.110	0.144	7.544	0	0.030
ALC	50.94	0.268	0.311	0.103	0.141	7.912	2	<b>0.117</b>
BAN	50.94	0.261	0.316	0.105	0.147	8.056	2	<b>0.157</b>
BGT	54.72	0.291	0.316	0.110	0.155	8.575	2	0.048
BLT	54.72	0.293	0.319	0.113	0.167	8.950	2	0.066
BLS	49.06	0.295	0.289	0.087	0.136	7.246	0	-0.040
BRI	52.83	0.276	0.298	0.105	0.146	7.879	2	0.053
CAT	50.94	0.275	0.309	0.100	0.141	7.674	2	0.076
CHW	50.94	0.284	0.292	0.097	0.137	6.867	0	-0.032
CIRP	49.06	0.261	0.311	0.097	0.137	7.681	0	<b>0.147</b>
DUB	56.60	0.254	0.277	0.102	0.136	7.434	2	0.027
EGE	50.94	0.254	0.286	0.094	0.131	7.198	1	<b>0.086</b>
GAE	52.83	0.295	0.319	0.111	0.151	8.347	2	0.042
GAR	52.83	0.251	0.305	0.102	0.144	7.938	1	<b>0.150</b>
GRK	52.83	0.263	0.297	0.100	0.151	8.045	1	0.102
GRM	50.94	0.263	0.314	0.104	0.144	7.890	2	<b>0.136</b>
HER	54.72	0.272	0.301	0.116	0.153	8.446	1	0.082
IMC	56.60	0.257	0.285	0.108	0.161	8.508	1	<b>0.097</b>
LAMW	52.83	0.265	0.293	0.095	0.144	7.633	1	0.067
LAS	52.83	0.261	0.301	0.101	0.148	8.028	1	<b>0.124</b>
MAK	50.94	0.288	0.308	0.101	0.151	7.903	1	0.040
MOM	52.83	0.252	0.290	0.094	0.137	7.622	0	<b>0.110</b>
NEA	52.83	0.291	0.317	0.114	0.160	8.556	1	0.066
ORAE	52.83	0.327	0.318	0.110	0.156	8.376	1	-0.060
ORAW	50.94	0.316	0.317	0.109	0.145	8.075	0	-0.025
ORB	50.94	0.268	0.300	0.102	0.141	7.523	1	0.076
PIST	54.72	0.255	0.278	0.095	0.146	7.755	1	0.065
PSL	50.94	0.284	0.322	0.101	0.154	8.301	1	0.111
SAR	50.94	0.276	0.302	0.104	0.143	7.741	1	0.067
SBRB	47.17	0.277	0.307	0.091	0.124	6.969	0	0.056
SET	50.94	0.296	0.327	0.119	0.151	8.161	1	0.064
SGL	52.83	0.276	0.316	0.108	0.162	8.564	1	<b>0.114</b>
SIS	52.83	0.247	0.295	0.103	0.147	7.840	1	<b>0.142</b>
TES	50.94	0.255	0.287	0.096	0.137	7.412	2	0.088
TURK	45.28	0.310	0.328	0.095	0.138	7.337	0	0.032
VAL	54.72	0.231	0.281	0.100	0.141	7.700	3	<b>0.161</b>
AZO	50.94	0.274	0.275	0.087	0.138	7.293	1	-0.006
LID	54.72	0.275	0.309	0.114	0.145	8.102	2	0.072
ORI	49.06	0.289	0.310	0.102	0.131	7.508	1	0.032
CAM	47.17	0.318	0.341	0.119	0.149	8.108	1	0.044
CAS	52.83	0.270	0.298	0.110	0.152	8.004	0	0.077
VIG	54.72	0.279	0.309	0.124	0.155	8.644	2	0.084
BID	50.94	0.319	0.338	0.122	0.158	8.713	1	0.035
CHA	39.62	0.257	0.281	0.075	0.112	5.954	2	0.061
LOI	49.06	0.224	0.244	0.083	0.114	5.978	0	0.061
MSMA	50.94	0.226	0.245	0.083	0.075	5.630	1	-0.007
KKAT	62.26	0.199	0.245	0.110	0.135	7.343	2	<b>0.146</b>

$P_O$ , % of polymorphic loci;  $H_O$ , observed heterozygosity;  $H_E$ , expected heterozygosity; MAF, minor allele frequency;  $F_{IS}$ , inbreeding coefficient; values with  $P < 0.05$  after Benjamini–Yekutieli correction are marked in bold.

populations from the Adriatic Sea region formed a separate, but not well supported, sub-group. The remaining *M. galloprovincialis* populations fell outside these two main clusters with little or no evidence of groupings based on geography.

Correspondence analysis (CA) of 43 *M. galloprovincialis* populations (of which 7 are reference populations) revealed clear separation of an Atlantic Ocean grouping, a Black Sea plus Sea of Azov grouping, and a third grouping consisting of all other populations, except the ORAW population that was located between the Atlantic Ocean and the Mediterranean Sea groups (Figure 3). After removal of Atlantic Ocean and African samples a CA carried out on 33 populations and including 3 reference populations (LID, ORI, AZO) resolved three groups: the *M. galloprovincialis* from the Black Sea and the Sea of Azov, *M. galloprovincialis* from the Mediterranean Sea, and *M. galloprovincialis* from the Aegean Sea (Figure 4).

STRUCTURE analysis of 46 *Mytilus* populations (i.e., excluding the reference *M. trossulus*) revealed that the largest increase of LnP(D) was for  $K = 2$  and then for  $K = 3$ . The greatest subdivision was detected at  $K = 2$  where the clusters corresponded to the separation of the *M. edulis* and *M. galloprovincialis* populations (Figure 5). At  $K = 3$ , the clusters corresponded to *M. edulis*, and two groups of *M. galloprovincialis*, with a general division into lineages from the Mediterranean Sea and the Atlantic Ocean because individuals are potentially admixed. Division between the *M. galloprovincialis* groups was most apparent at the population level because the frequency of one of the clusters characteristic of *M. galloprovincialis* was >77% (average 80%) in Atlantic Ocean populations, whereas the frequency of the second cluster was >50% (average 61%) in *M. galloprovincialis* from the Mediterranean Sea populations (Fig. S3). From the STRUCTURE analysis when  $K = 3$  most of the 1292 individuals (97.2%) were correctly assigned to their original sample taxon (one of the three clusters) with  $q > 0.8$ . Mussels collected from Morocco (AGA, Atlantic Ocean) and Algeria (ORAW, Mediterranean Sea) showed gene admixture characteristic of *M. galloprovincialis* reference groups from the Atlantic Ocean (BID, VIG, CAS and CAM). A second sample from Algeria (ORAE), located east of the Alboran Front, showed gene admixture characteristic of *M. galloprovincialis* populations from the Mediterranean Sea.

#### *Linking genetic and environmental variation*

The GLM testing of the “All effects” model was significant ( $P < 0.05$ ), with Mixed Layer thickness ( $P < 0.0001$ ), PAR at surface ( $P < 0.004$ ), PAR on

seafloor ( $P < 0.011$ ), Cloud cover ( $P < 0.03$ ) and SST ( $P < 0.036$ ) as significant terms. The other eight environmental variables did not explain significant variation in the population-specific  $F_{ST}$  values ( $P > 0.05$ ). For the DistLM analysis the best-fit models built using adjusted  $R^2$ , AIC and BIC gave the same results so we report here only the results for AIC. The marginal test results (Table V) indicated that seven of the 13 environmental variables explained significant variation in population-specific SNP locus allele frequencies. These seven variables explained a total of 75% of the variation in the SNP dataset. In the sequential testing the best-fit model contained all 11 environmental variables, of which five variables were statistically significant (Wave Ht, PO<sub>4</sub>, SST, PAR at surface, Mixed layer) in a 13-term model that explained 61.4% of the variation in the SNP dataset (Table V).

## Discussion

The Mediterranean Sea is characterised by a high biodiversity of marine organisms and a large number of endemic species (Coll et al. 2010; Danovaro et al. 2010). Despite the connection with the Atlantic Ocean via the Strait of Gibraltar and the possibility of transport of planktonic larvae and motile adults, many species show genetic distinctiveness of their Mediterranean Sea populations (Patarnello et al. 2007; Pascual et al. 2017). These species include the cirriped *Chthamalus montagui* (Shemesh et al. 2009; Pannacciulli et al. 2017), the shrimp *Palaeomon elegans* (Reuschel et al. 2010) and the blue mussel *Mytilus galloprovincialis* (Varvio et al. 1988; Quesada et al. 1995b; Kijewski et al. 2011; Zbawicka et al. 2012; Del Rio-Lavín et al. 2022).

The location of the Almeria-Oran Front is the area where two circular sea currents meet and flow from north to south-east into the Mediterranean Sea and then south-west, which directs water of Atlantic Ocean origin into the Alboran Sea. This is considered to be an isolating factor (Millot & Taupier-Letage 2005; Millot 2013; Pascual et al. 2017) for some marine species between the Atlantic Ocean proper and the Mediterranean Sea. Abrupt changes of allele frequencies in *M. galloprovincialis* populations related to this front have been reported for a range of difference genetic marker types, including allozymes (Sanjuan et al. 1994; Quesada et al. 1995a), mtDNA (Quesada et al. 1995b), microsatellites (Diz & Presa 2008; Ouagajjou et al. 2010; Ouagajjou & Presa 2015) and a few nuclear DNA markers (El Ayari et al. 2019). In addition to environmental factors, El Ayari et al. (2019) also highlight the possibility of intrinsic mechanisms, such as

Table IV. Analysis of molecular variance (AMOVA) calculated for five scenarios (different groupings of samples). All values were significant for a P value &lt;0.05.

Scenario	Among groups			Among populations within groups			Among individuals within populations			Within individuals		
	Variance component	% variation	Variance component	% variation	Variance component	% variation	Variance component	% variation	Variance component	% variation	Variance component	% variation
1 - Mediterranean Sea only /1 group/	0.83	20.86	—	—	0.28	0.28	6.98	2.89	72.16	72.16	—	—
2 - Mediterranean Sea only /2 groups/	0.11	3.20	0.02	0.52	0.30	9.07	2.90	87.21	87.21	—	—	—
3 - Mediterranean Sea only /4 groups/	0.04	1.23	0.01	0.43	0.29	9.02	2.91	89.33	89.33	—	—	—
4 - Atlantic Ocean & Mediterranean Sea incl. ORAW	0.07	1.64	0.82	20.25	0.28	6.89	2.89	71.22	71.22	—	—	—
5 - Atlantic Ocean incl. ORAW & Mediterranean Sea	0.06	1.50	0.82	20.26	0.28	6.90	2.89	71.34	71.34	—	—	—

**Scenario 1 - Mediterranean Sea only /1 group/** = 38 populations (pooled ALC, BAN, BGT, BLT, IMC, LAMW, LAS, ORAE, PSL, SET, SGL, VAL, ORI, ORAW, ORB, GAE, NEA, PIST, CIRP, BRI, GAR, CAT, LID, SIS, CHW, MAK, DUB, GRK, GRM, SAR, TES, EGE, TURK, HER, SRB, AZO, BLs, MOM).

**Scenario 2 - Mediterranean Sea only /2 groups/** = Mediterranean Sea (pooled ALC, BAN, BGT, BLT, IMC, LAMW, LAS, ORAE, PSL, SET, SGL, VAL, ORI, ORAW, ORB, GAE, NEA, PIST, CIRP, BRI, GAR, CAT, LID, SIS, CHW, MAK, DUB, GRK, GRM, SAR, TES, EGE, TURK, HER); Black Sea and Sea of Azov (pooled SRB, AZO, BLs, MOM).

**Scenario 3 - Mediterranean Sea only /4 groups/** = Western Mediterranean Sea (pooled ALC, BAN, BGT, BLT, IMC, LAMW, LAS, ORAE, PSL, SET, SGL, VAL, ORI, ORAW); Central Mediterranean Sea (pooled ORB, GAE, NEA, PIST, CIRP, BRI, GAR, CAT, LID, SIS, CHW, MAK, DUB, GRK, GRM); Aegean Sea (pooled SAR, TES, EGE, MOM, TURK, HER); Black Sea and Sea of Azov (SRB, AZO, BLs).

**Scenario 4 - Atlantic Ocean & Mediterranean Sea incl. ORAW** = Atlantic Ocean (pooled CAS, BID, VIG, CAM, AGA); Mediterranean Sea and Black Sea and Sea of Azov (pooled ALC, BAN, BGT, BLT, IMC, LAMW, LAS, ORAE, PSL, SET, SGL, VAL, ORI, ORAW, ORB, GAE, NEA, PIST, CIRP, BRI, GAR, CAT, LID, SIS, CHW, MAK, DUB, GRK, GRM, SAR, TES, EGE, TURK, HER, SRB, AZO, BLs, MOM).

**Scenario 5 - Atlantic Ocean incl. ORAW & Mediterranean Sea** = Atlantic Ocean (pooled CAS, BID, VIG, CAM, AGA, ORAW); Mediterranean Sea and Black Sea and Sea of Azov (pooled ALC, BAN, BGT, BLT, IMC, LAMW, LAS, ORAE, PSL, SET, SGL, VAL, ORI, ORB, GAE, NEA, PIST, CIRP, BRI, GAR, CAT, LID, SIS, CHW, MAK, DUB, GRK, GRM, SAR, TES, EGE, TURK, HER, SRB, AZO, BLs, MOM).

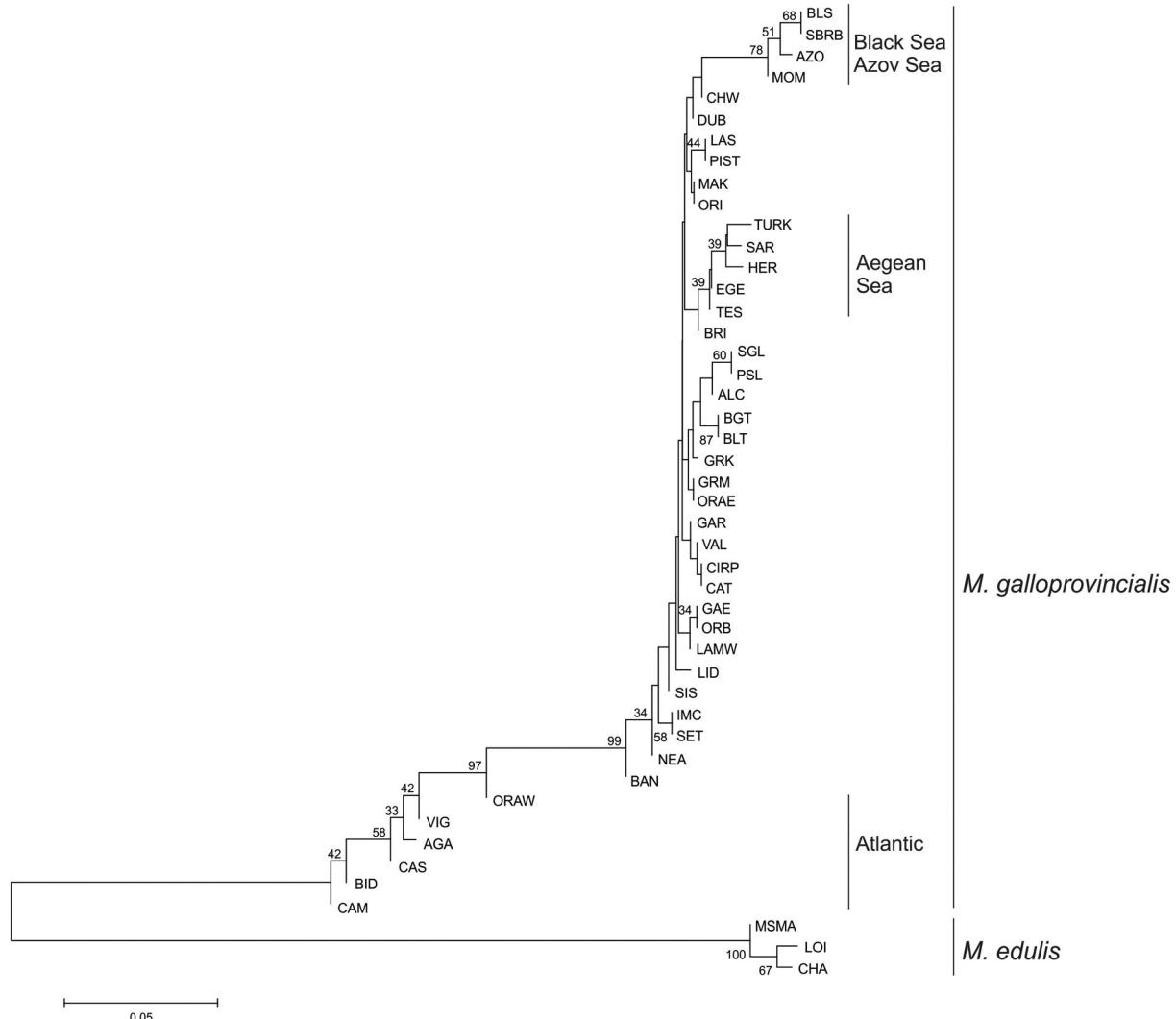


Figure 2. Neighbour-joining tree of *Mytilus galloprovincialis* populations from the Mediterranean Sea, Black Sea and the Sea of Azov, and reference populations of *M. edulis* and *M. galloprovincialis* based on the  $F_{ST}$  distance measures obtained with POPTREEW and visualised with MEGA version 6. The numbers at the nodes represent percent bootstrap values estimated from 10,000 replicates (only values above 30 are shown).

pre- or post-zygotic isolation, in the maintenance of the blue mussel hybrid zone on the Algerian coast of the Mediterranean Sea, just east of Oran. Our SNP results confirm the occurrence of the Atlantic Ocean lineage of *M. galloprovincialis* on the Alboran African coast (sample ORAW) and the Mediterranean Sea lineage east of Oran (ORAE).

In the present study, significant differentiation between mussel populations of the Sea of Azov and the Black Sea from the Mediterranean Sea populations was identified with two of the 53 SNP loci. No statistically significant differences were observed among populations from the Black Sea and the Sea of Azov and the Dardanelle Straits. Most genotypes were common to the Mediterranean, Black and

Azov Seas and what differences existed are attributable to genotype frequency differences among the populations. A few SNP loci with alleles rare in the Black Sea and Sea of Azov populations were more frequent in the Mediterranean Sea populations (e.g., BM102A “C”, BM105A “G”, BM26B “T”), or those that did not occur in the Black Sea and Sea of Azov, for example, BM106B “G”. Only two genotypes were more frequent in the Sea of Azov than in Black Sea populations: BM147A “C” and BM32A “A”. Significant differentiation between mussel populations from the Mediterranean and Black Seas has also been found using as many as 512 SNP loci (Paterno et al. 2019) although no significant genetic structuring was noted within the

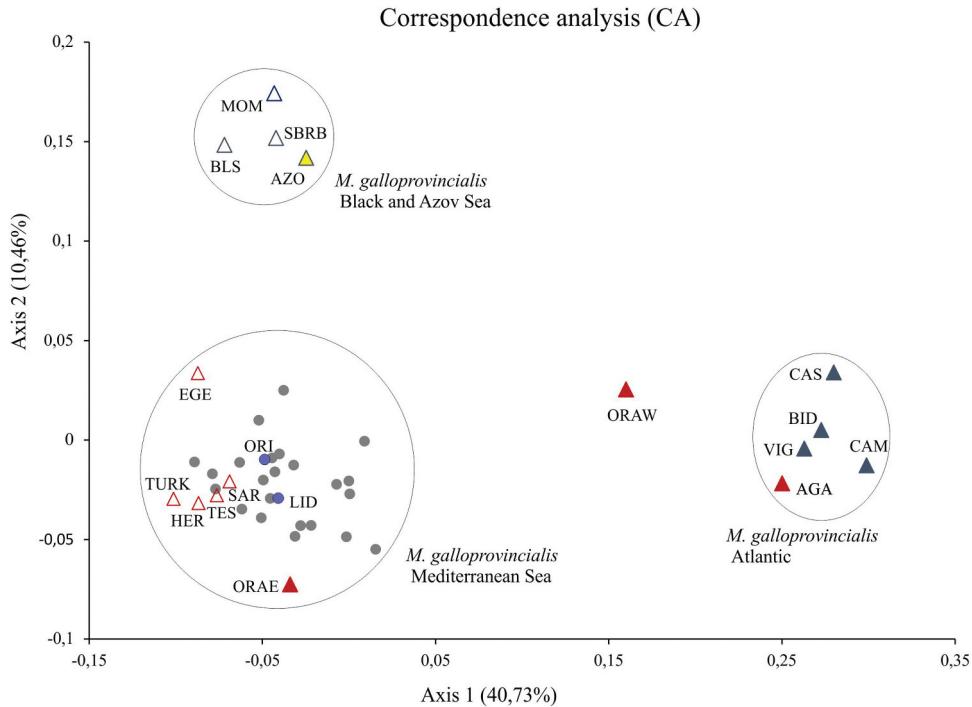


Figure 3. The first two axes of the correspondence analysis (CA) computed from the SNP locus variation among 36 *M. galloprovincialis* populations from the Mediterranean Sea, Black Sea and Atlantic Ocean and seven reference populations of *M. galloprovincialis*). Samples from Morocco (AGA - Atlantic Ocean) and Algeria (ORAE and ORAW - Mediterranean Sea) are shown as red filled triangles. Samples from the Aegean Sea region (EGE, HER, SAR, TES, Turk) are shown as red unfilled triangles. Each point or triangle is one of the studied populations. LID and ORI (blue filled triangles), AZO (yellow filled triangle) and BID, CAM, CAS and VIG (grey filled triangles) are reference samples.

Black Sea populations using genotyping of 998 SNP loci (Paterno et al. 2019). Genetic divergence has also been reported for *M. galloprovincialis* from Mediterranean Sea and Black Sea populations at the mtDNA level (Ladoukakis et al. 2002). RFLP analysis of COIII gene polymorphism in *M. galloprovincialis* populations revealed that almost all Black Sea haplotypes also occur in the Mediterranean Sea populations, that is, the Black Sea populations are a sub-set of the Mediterranean Sea populations. In addition, no significant differences in allele frequencies have been reported between populations of *M. galloprovincialis* from the southern Black Sea, Bosphorus Strait and Sea of Marmara by assessment of variation in the COIII mtDNA region and six microsatellite loci (Kalkan et al. 2011). It has been suggested that a shift from Black Sea-like to Mediterranean Sea-like genetic structure occurs at a location in the Dardanelle Strait, which is supported by our SNP data. The most common mtDNA haplotype identified by RFLP analysis of ND2-COIII in a population from the Sea of Azov was also present in a sample from Villafranche-sur-Mer, northwestern Mediterranean Sea (Śmiertanka et al. 2004). However, a few rare

haplotypes were unique. Genotyping-by-sequencing of the same region of mtDNA revealed haplotype frequency differences between populations from the Sea of Azov and Black Sea, but most common haplotypes were also present in the northwestern Mediterranean Sea (Gerona, Banyuls sur Mer and Gulf of Oristano) (Śmiertanka et al. 2009, 2014).

Because the timing of onset of the flow of Mediterranean Sea waters into and possible colonisation by euryhaline bivalve populations of the Black Sea has not been defined precisely (Nikula & Väinölä 2003; Sromek et al. 2019), we have reviewed in detail here the information available. It is estimated that Mediterranean Sea water from the Aegean Sea filled the Sea of Marmara and replaced freshwater originating from the Black Sea approximately 12 K ybp (Algan et al. 2001). The outflow of lacustrine water from the present area of the Black Sea probably occurred between 10 to 8.4 K ybp, which was followed by pulses of Mediterranean Sea seawater entering the Black Sea (Hiscott et al. 2007). Connection of the Black Sea to the Mediterranean Sea may have taken place 9.4 k ybp, as indicated by strontium/oxygen ratio changes in molluscan shells (Major et al. 2006). Transition

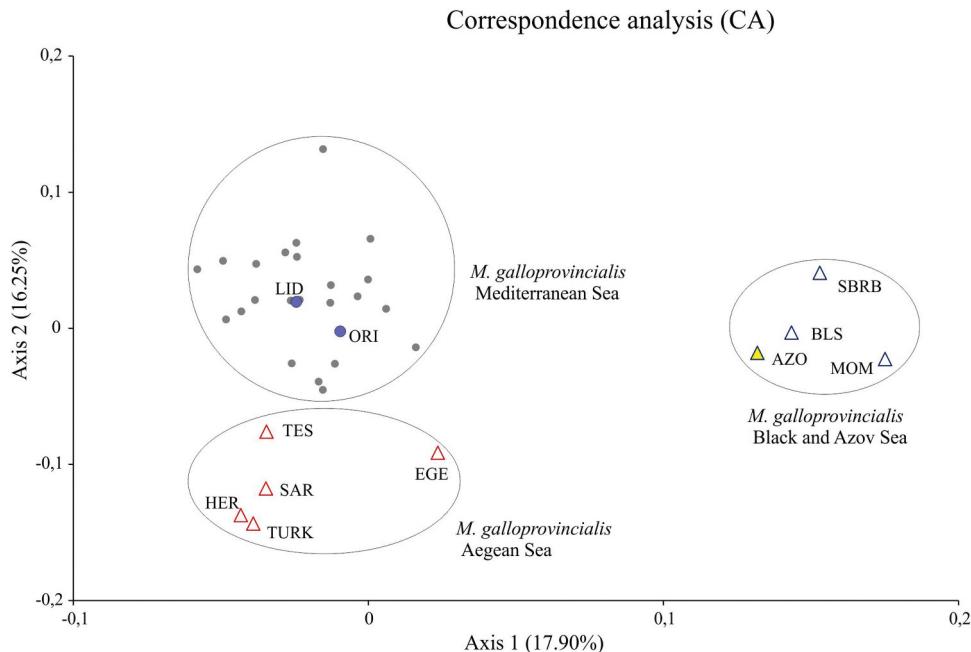


Figure 4. Correspondence analysis (CA) plot computed from the SNP locus variation among 34 populations of *M. galloprovincialis* from the Mediterranean Sea, Aegean Sea and Black Sea. Each point or triangle is one of the studied populations. LID and ORI (blue filled triangles) and AZO (yellow filled triangle) are reference samples.

of the Black Sea from a freshwater lake to a marine environment most probably occurred between 8.9 and 8.5 K ybp (Tyuleneva et al. 2014). Outflow of freshwater or isolation lasted until 8 K ybp (Soulet et al. 2011) and was followed by establishment of today's hydrographic system with surface outflow of low salinity Black Sea water and inflow at depth of Mediterranean Sea high salinity waters (Kokkos & Sylaios 2016) via the Bosphorus Strait (Hiscott et al. 2007). The first representatives of the Mediterranean mollusc fauna appeared in deposits in the Bosphorus region 5.3 K ybp, whilst *Mytilus* sp. appeared for the first time ~4.4 K ybp (Algan et al. 2001). In a core from the southwest Black Sea, *M. galloprovincialis* has been found in layers dated to 5.9–2.4 K ybp (Hiscott et al. 2007). Filipova-Marinova et al. (2013) observed the occurrence of a *M. galloprovincialis* shell layer in deposits of Varna Lake (Bulgarian Black Sea) that were radiocarbon dated to 7,776 to 6,183 ybp. According to Tyuleneva et al. (2014) timing of the arrival of *M. galloprovincialis* in the northwestern Black Sea (estimated without  $^{14}\text{C}$  dating) was from the Bugazian at 10.5–8.4 K ybp, Kalamitan beds at 7.1 to 4 K ybp and Dzhemetician beds at 4.1 K ybp to the present in sediments containing marine euryhaline species. A discrepancy in time estimation of the earliest appearance of *Mytilus* in Black Sea deposits may be

related to elution and displacement of some sediment layers. It can be assumed that *Mytilus* populations settled the Black Sea sometime since 8 K ybp as a result of post-glacial expansion from refugia.

Time of divergence of mtDNA haplotypes in the Black Sea and Mediterranean Sea mussel populations has been estimated as a few hundred thousand years ago (Śmietanka et al. 2014), which is much earlier than the origin of the mussel population in the Black Sea. Therefore, euryhaline *M. galloprovincialis* populations evolved in the Mediterranean Sea and the Aegean Sea (possibly including the Levantine Sea) long before the contemporary marine phase of the Black Sea. Similar to the Black Sea situation, the Baltic Sea in northern Europe was a freshwater lake and became connected with the Atlantic Ocean and newly developing North Sea approximately 8–9 ybp (Berglund et al. 2005; Behre 2007). For comparison, Kostecki and Janczak-Kostecka (2011) have reported the onset of the marine water environment in Pomeranian Bay, southwest Baltic Sea, to be 8.9–8.3 ybp as determined using geochemical estimators. The Baltic Sea is characterised by lower marine species diversity in comparison to the present-day North Sea (Johannesson & Andre 2006; Wennerström et al. 2017). Genetic divergence of *Mytilus* populations in both regions (i.e., the Black Sea from the

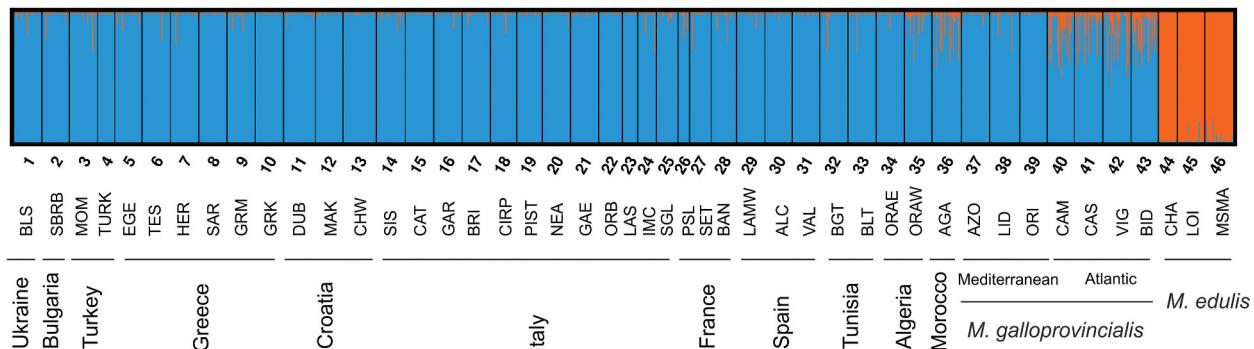
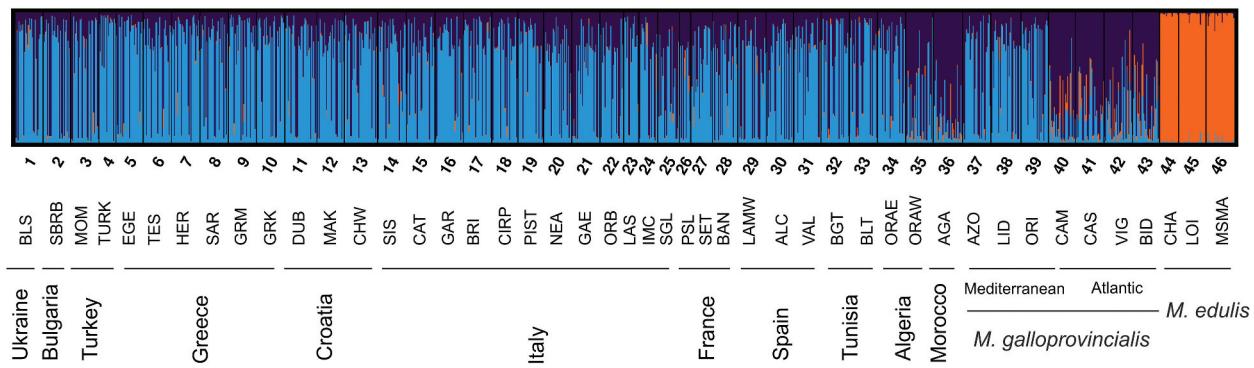
$K = 2$  $K = 3$ 

Figure 5. Plot from STRUCTURE analysis for 46 populations (i.e., excluding reference KKAT *M. trossulus*) at  $K = 2$  and  $K = 3$ . Each individual is represented by a single vertical line broken into two or three coloured segments, with lengths proportional to each of the  $K$  inferred clusters. Site abbreviations of the samples are provided in Table I. Vertical black lines separate the samples.

Mediterranean Sea and the Baltic Sea from the North Sea) is recent in evolutionary time and can be related to a bottleneck effect, subsequent expansion of populations and selection factors acting via brackish water conditions. A characteristic of the *M. edulis* and *M. trossulus* SNP allele BM105A “A” is its increased frequency among mussels in the Black Sea and Sea of Azov and its decreased frequency among mussels in the Mediterranean Sea. Nevertheless, the absence of *M. edulis* and *M. trossulus* genotypes, as determined from our studies and other published research, excludes the possibility that *Mytilus* populations from areas such as the White Sea and/or the Barents Sea have invaded the Black Sea from north-east European seas via routes now closed to them.

Mitochondrial DNA genetic analyses of *M. galloprovincialis* from northern, central and southern Greek coasts of the Aegean Sea revealed that these populations were homogeneous (Ladoukakis et al. 2002; Giantsis et al. 2014b). Moreover, Giantsis et al. (2014b) reported that the population from the area closest to the Dardanelle Strait (Sea of

Marmara, Turkey) differed significantly from the Aegean Sea populations, whereas differences between samples from the Aegean, Ionian and North Adriatic Seas were weak with the exception of one sample from the area of Zadar (Croatia). MtDNA sequence analyses revealed genetic homogeneity among all Greek populations and the clear differentiation of the only Turkish sample (Çanakkale, Dardanelle Strait, Sea of Marmara) from the Aegean populations (Giantsis et al. 2014b). On the other hand, microsatellite data suggested significant differentiation of Italian samples (from the North Adriatic Sea and Ligurian Sea, respectively) from Aegean samples (Giantsis et al. 2014a). Genotyping with a large number of SNPs in a more local study revealed population variation in the northern and southern Adriatic Sea different from that in the northern Ionian Sea (Paterno et al. 2019). The results of the present study are in agreement with the findings of these different studies and extend the microsatellite DNA analyses (Giantsis et al. 2014a) concerning the genetic

Table V. Marginal (top) and sequential (bottom) test results from the DistLM analysis.

Marginal tests			P	% Prop.
Variable	SS(trace)	Pseudo-F	P	% Prop.
Wave Ht	5.312	15.586	0.0001	27.54
PO <sub>4</sub>	1.952	4.617	0.0090	10.12
SST	1.919	4.531	0.0037	9.95
PSU	1.800	4.210	0.0113	9.31
2 m air temp	1.226	2.782	0.0307	6.36
Cloud cover	1.182	2.677	0.0343	6.10
PAR at surface	1.162	2.628	0.0342	6.02
10 m wind	0.993	2.225	0.0642	5.15
NO <sub>3</sub>	0.850	1.890	0.0793	4.41
Prim Prodn	0.794	1.761	0.1051	4.12
Mixed layer	0.714	1.577	0.1446	3.70
Chl	0.513	1.120	0.2972	2.66
PAR at seafloor	0.348	0.754	0.5503	1.81

Sequential tests			P	%Prop.	Cumul. prop
Variable	AIC	SS(trace)	Pseudo-F	P	%Prop.
+Wave Ht	-44.332	5.31	15.586	0.0001	27.54
+PO <sub>4</sub>	-48.595	1.89	6.272	0.0001	9.82
+SST	-51.451	1.29	4.662	0.0004	6.69
+PSU	-50.008	0.14	0.496	0.9056	7.21
+2 m air temp	-48.660	0.16	0.565	0.8606	8.30
+Cloud cover	-48.035	0.33	1.170	0.2807	1.71
+PAR at surface	-50.587	1.02	3.908	0.0011	5.29
+10 m wind	-49.595	0.21	0.806	0.6161	1.10
+NO <sub>3</sub>	-49.843	0.45	1.772	0.0705	2.36
+Prim Prodn	-48.824	0.19	0.738	0.6907	0.99
+Mixed layer	-49.726	0.54	2.164	0.0258	2.80
+Chl	-48.832	0.20	0.782	0.6374	1.02
+PAR at seafloor	-47.431	0.10	0.407	0.9690	0.05

+ indicates the addition of a given term into the full model, variables being added in the order shown.  
 Variable name (refer to Table S1): 2 m air temp - 2 metre air temperature; 10 m wind - 10 metre wind speed; Chl - Total chlorophyll; Cloud cover - Total cloud cover; Mixed layer thickness; NO<sub>3</sub> - Nitrate concentration; PAR at seafloor - Photosynthetically available radiation on the floor of the ocean; PAR at surface - Photosynthetically available radiation at ocean surface; PO<sub>4</sub> - Phosphate concentration; Prim Prodn - Phytoplankton primary production; PSU - Salinity; SST - Sea surface temperature; Wave Ht - Significant height of combined wind waves and swell.

structure of East-Central Mediterranean *M. galloprovincialis* populations. No differences between northern and southern Adriatic Sea *M. galloprovincialis* populations were observed in our study by genotyping with 53 SNPs. However, the Çanakkale (Sea of Marmara) population differed from the Aegean Sea samples and the Çanakkale sample is representative of the Black Sea mussel population genetic variation. A very rare allele (BM5B "G") was more frequent in Aegean Sea mussels than in all other *M. galloprovincialis*, *M. edulis* and *M. trossulus* populations. Such subdivision of populations within the Eastern versus Western basins of the Mediterranean Sea has also been reported for other species, for example, the marbled goby *Pomatoschistus marmoratus* (Mejri et al. 2011), the sea star *Astropecten aranciacus* (Zulliger et al. 2009) and the cockle *Cerastoderma glaucum* (Nikula & Väinölä 2003).

We did not observe significant genetic differentiation among populations of *M. galloprovincialis* in the Western Mediterranean Basin, east of the Alboran Front to the coasts of Italy and Tunisia. Similarly, analysis of samples from Banyuls (France) and Haouaria (Tunisia) using 512 SNP loci did not identify differences (Paterno et al. 2019). Lack of genetic differentiation among *M. galloprovincialis* populations within the Western Basin can be partly explained by natural reasons, such as extensive gene flow and a common paleogeographic history of this Basin including the Balearic, Ligurian and Tyrrenian Seas separated from the Eastern Basin by the Siculo-Tunisian Strait. After each glacial phase since the Pleistocene (cycles of 41 K years prior to and 100 K years after the Mid Pleistocene Transition), *M. galloprovincialis* populations expanded into newly flooded areas as a result of sea level rising and then retreated with sea level falling (repeated interglacial colonisation and expansion, and retreat during glaciations). However, the biggest drop in sea level occurred 24–19 K ybp during the Last Glacial Maximum (Boavida et al. 2019). Surface waters in the central Mediterranean Sea (in the vicinity of the Siculo-Tunisian Strait) were most probably much colder in spring during glacial periods in comparison to interglacial periods (Rouis-Zargouni et al. 2010), which suggests much colder winter temperatures in comparison to the present. Strong meltwater discharge also probably influenced cooling (lower temperatures) and increased variability of the salinity of coastal Mediterranean Sea waters. Increase in rainfall during the past interglacial/glaciation periods resulted in a decrease of sea surface salinities e.g., in the northern Tyrrhenian Sea (Dixit et al. 2020) and increased

suspended matter concentrations, and intensified the flux of nutrient and organic matter in the Mediterranean Sea causing deposition of sapropels (Toucanne et al. 2015) in both the Eastern and Western basins (Rohling et al. 2015). Thus, a combination of physico-chemical changes in sea water composition may explain genetic differentiation observed in contemporary Western and Eastern Mediterranean Sea populations of *M. galloprovincialis*. A separation of western and eastern Mediterranean populations by a possible barrier of periodically shallow water in the central Mediterranean Sea (the Siculo-Tunisian Strait) has been indicated by Chefaoui et al. (2017). A divergence between western and eastern Mediterranean Sea populations has also been found for species such as the sea bass *Dicentrarchus labrax* (Bahri-Sfar et al. 2000), the sea cucumber *Holothuria polii* (Valente et al. 2015), the crab *Carcinus aestuarii* (Ragionieri & Schubart 2013), the hermit crab *Diogenes pugilator* (Almón et al. 2021), and the cockle *Cerastoderma glaucum* (Sromek et al. 2019).

Seascape genetics seeks to identify associations between environmental variation and genetic variation with the ultimate aim of identifying key environmental factors that contribute to explanation of population genetic variation and regional differences in genetic structure (e.g., Selkoe et al. 2008, 2016; Riginos & Liggins 2013; Wei et al. 2013; Silva & Gardner 2016; Zeng et al. 2020). Typically this approach, which is most often based on variation of neutral genetic markers, seeks to identify environmental factors that promote (e.g., currents) or retard (e.g., salinity variation acting as a barrier) gene flow. Whilst this new approach to understanding connectivity is both powerful and elegant, it is often limited by data availability - the low numbers of environmental variables that are available for multiple sites within a region of study. In the present study we used 13 environmental variables collated from 43 sites within the Mediterranean Sea. The GLM analysis identified five environmental variables (mixed layer thickness, PAR at surface, PAR at seafloor, SST, Cloud cover) that explained variation in population-specific  $F_{ST}$  values. In contrast, the sequential tests for the DistLM analysis best-fit model, which explained 61% of the variation in the raw SNP data set, included all 13 environmental variables, although only four variables (Wave height, PO<sub>4</sub>, SST, PAR at surface) were significant in the multi-term model. Earlier literature suggests that salinity and temperature often drive the generic patterns of hard-bottom intertidal species including *Mytilus* spp. (Kaiser et al. 2011) and this result was observed for SST but not for salinity (PSU) in the *M.*

*galloprovincialis* of the Mediterranean Sea. The absence of salinity from our significant results may reflect the fact that the gradient in the Mediterranean Sea is not great, and certainly not as great as in other regions (e.g., in the Baltic Sea – Kijewski et al. 2019). Overall, our results suggest that there is a complex mix of environmental variables that contribute to genetic variation of *M. galloprovincialis* populations in the Mediterranean Sea, rather than a simple (one or two variable) explanation as has been reported for several other coastal and deep-sea marine invertebrates (e.g., Wei et al. 2013; Silva & Gardner 2016; Zeng et al. 2020). The complexity of the seascape genetics results in the present study may reflect (1) the reasonably large number of environmental variables in our data set ( $n = 13$ ) and the fact that many different environmental variables are likely to influence gene flow in any given system (i.e., we might expect to detect a complex result simply because we have a complex environmental data set) and/or (2) the complex geological history of the Mediterranean Sea and the formation of its waters and the associated sub-basins of the region leading to complex interactions between environmental and genetic variation that may be site-specific or regional rather than basin-wide. In addition, as noted above, seascape genetic analyses usually focus on detecting associations across neutral loci. Many of the SNPs loci employed here are in coding regions even if they do not exhibit outlier status (refer to Figs. S1 and S2). Thus, the detection of an association between genes under (low) selective pressure and environmental variation cannot be ruled out. Elsewhere, Sun and Hedgecock (2017) have highlighted for high gene-flow species (*M. galloprovincialis* falls into this group) the need to better understand the role that temporal genetic variation may play in contributing to seascape genetics analysis results from a single (snap-shot) version to a temporal sampling series. Finally, it is worth noting that the GLM and DistLM analyses are both linear-based methods that do not include interaction terms. Other analytical approaches, such as boosted regression trees (Elith et al. 2006, 2008; Leathwick et al. 2006; Hastie et al. 2009; Kotta et al. 2017) may prove to be more informative because they are not constrained to detect only linear relationships and can examine interactions among variables, but they require larger (training) data sets than are often available to most researchers and to us for the present study.

The contemporary genetic structure of the Mediterranean Sea populations of *M. galloprovincialis* is the result of a combination of natural and anthropogenic factors. The dispersal ability of the species is

expected to result in high gene flow and connectivity among populations and lead to genetic homogenisation on a broad spatial scale. Although this could explain the homogenisation observed within individual basins, it cannot explain the comparative homogenisation found among the different basins. For instance, based on the topography and oceanographic conditions of the Siculo-Tunisian Strait in the central Mediterranean Sea, one would expect to observe (at least a degree of) genetic differentiation between samples from the Adriatic Sea and the Ligurian Sea, which this study did not find. Anthropogenic activities, including hull fouling, transport of ballast water, the movements of exploration or drilling rigs and unrecorded human-mediated transplantation of spat for aquaculture may all have played an important role in overcoming natural barriers to mussel connectivity and consequently may have contributed to the shaping of the present patterns of genetic homogeneity among the Central-Western Mediterranean Sea mussel populations. For example, the *M. galloprovincialis* Atlantic form has been reported to have been introduced via ballast water and/or hull fouling to ports in an area from the north of France to Norway, where it has hybridised with local *M. edulis* (Simon et al. 2020). These hybrids have been termed “dock mussels” (because they are mostly found in docks associated with shipping activity) and have appeared in recent decades rather than centuries. In the opposite direction, Mediterranean ports may also experience strong pressure from invasive species, including *M. edulis*. For example, *M. edulis* specimens have been identified using molecular genotyping on a barge hull, which arrived from Middlesborough, northern England, and was moored near the wreck of the cruise vessel *Costa Concordia* (Isola del Giglio, north Tyrrhenian Sea) in 2012 (Casoli et al. 2016). However, these mussels experienced a high mortality rate at temperatures over 22°C and did not live long enough to interbreed with local populations of *M. galloprovincialis*. Prior to this, Beaumont et al. (2006) reported anecdotal evidence of the introduction of *M. edulis* to the Mediterranean coast of France for aquaculture.

Development of an aquaculture industry targeted on *Mytilus* spp. production and other commercially important shellfish in the twentieth century contributed to unintentional introductions and the spread of many species (Coll et al. 2010) and can also be considered as a vector of mussel transportation between basins in the Mediterranean Sea. An excellent example is the case of Chalastra, a culture area near Thessaloniki, Greece, which serves as the main mussel culture area in the Aegean Sea. Mussel spat from this area have been repeatedly translocated to other mussel farms throughout Greece, Italy and France for ongrowing. Given that these



Figure 6. Rafting on anthropogenic plastic objects: detached mussel farm buoy on the beach near Gargano (GAR), Adriatic Sea in 2012. Photo by Marta Póltorak.

translocations are unrecorded it is very difficult to estimate the magnitude of their contribution to the current lack of geographical structure observed within Mediterranean Sea *M. galloprovincialis* populations. An example is the similarity of the Adriatic sample (CHW) to the Black Sea population (BLS).

Growing attention has been paid to the spread of *Mytilus* spp. by rafting on natural or artificial floating objects, including anthropogenic litter (Miller et al. 2018; Rech et al. 2018; Zbawicka et al. 2019; reviewed by Gardner et al. 2021). Plastic debris has been indicated as a transport vector for *Mytilus* sp./spp. on long distances in the North Atlantic and Arctic Oceans (Kotwicki et al. 2021). The main contributor to this spread via rafting is mariculture-related gear, including detached buoys (Figure 6). Interestingly, *M. edulis* that originated from market discards, aquaria discharged into the sea or introduced for aquaculture purposes in Venezia, Italy (Crocetta 2012), was not observed in our samples and most probably did not establish populations on the coast of Italy. In addition, our results do not confirm the existence of *M. edulis* (the so-called rock mussel grown in the Ebro Delta for aquaculture purposes, Wenne et al. (2022)) among native *M. galloprovincialis* populations on the coasts of Spain in the Balearic Sea.

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No potential conflict of interest was reported by the author(s).

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