1	Branched GDGT variability in sediments and soils from
2	catchments with marked temperature seasonality
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24 Abstract

25 The distributions of branched glycerol dialkyl glycerol tetraethers (brGDGTs) in sediments are used as 26 a proxy measurement to infer changes in past mean annual air temperatures (MAT). When applied in 27 high resolution sedimentary sequences, measurement of brGDGT distributions is employed to 28 reconstruct MAT at subdecadal time scales. In addition, brGDGT proxy estimates are also sometimes 29 purported to be seasonally biased in environments where annual brGDGT production may not be 30 constant during a seasonal cycle. The main aim of this study was to assess the occurrence of seasonality 31 in the production and distribution of brGDGTs, and the sesonality bias of the derived temperature proxy. 32 For this purpose, we examined brGDGT distributions and brGDGTs-derived MAT estimates, in surface 33 soils and settling/suspended particulate matter over one year from two sites located in the same latitude 34 but at different altitudes, in the Catalan Pyrenees, as well as at one site in southern Norway. These 35 locations have marked seasonal temperature cycles, which were expected to maximize the possibility of 36 detecting any seasonal bias in the production and compositions of brGDGTs. The results show that 37 brGDGT abundance is heterogeneous and increases with soil humidity. The brGDGT distributions and some of the brGDGT-derived proxy measurements in soils are relatively stable throughout the year and 38 39 do not change significantly in the suspended particulate matter in the river or settling particulate matter 40 in traps. Our study shows that the impact of the seasonality of temperature on brGDGT distribution was 41 absent in the soils studied, regardless of altitude or latitude on a catchment/regional scale. As soils are likely to contain a brGDGT signature which is representative of average environmental conditions in 42 43 the catchment at least over decades, brGDGT proxy reconstructions derived from soil sources are more 44 suitable to infer variability in environmental parameters over the same timescales (i.e. decades or longer). On shorter timescales (i.e. annual), sediment downcore variability in brGDGTs is likely to be 45 46 related to changes derived from *in situ* production and sediment sources.

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48 1. Introduction

Branched glycerol dialkyl glycerol tetraethers (brGDGTs; Fig. S1), which contain 4–6 methyl groups and 0-2 cyclopentane rings, are common lipids in terrestrial and aquatic environments. They are thought to originate from heterotrophic bacteria (Weijers et al., 2006, 2010; Sinninghe Damsté et al., 2011) that preferentially grow under suboxic conditions, possibly as facultative aerobic heterotrophs (Ayari et al., 2013; Huguet et al., 2013a; Huguet et al., 2017). However, their phylogeny and habitat diversity remains to be established.

55 BrGDGTs have been described in peat bogs (Schouten et al., 2000; Weijers et al., 2006; Huguet et al., 2010a, 2013b; Naafs et al., 2017a), soils (Weijers et al., 2007; Sinninghe Damsté et al., 2008; Bendle 56 57 et al., 2010; Huguet et al., 2010b; Loomis et al., 2011; Menges et al., 2014; Yang et al., 2014a; Warden 58 et al., 2016; Coffinet et al., 2017), stalagmites (Yang et al., 2011; Blyth and Schouten, 2013), water 59 suspended particulate matter (SPM) in continental settings (Blaga et al., 2009, 2011; Bechtel et al., 2010; Fietz et al., 2012; Schoon et al., 2013; Buckles et al., 2014; De Jonge et al., 2014a; Loomis et al., 2014; 60 Hu et al., 2016), river and lake sediments (Blaga et al., 2009; Tierney and Russell, 2009; Zink et al., 61 62 2010; Loomis et al., 2011; Sun et al., 2011; Zhang et al., 2012; Shanahan et al., 2013; Ajioka et al., 2014a; Zell et al., 2014a; Hanna et al., 2016; Freymond et al., 2017; Peterse and Eglinton, 2017), 63 64 windborne particulate matter (Fietz et al., 2013; Weijers et al., 2014), marine coastal (Hopmans et al., 2004; Kim et al., 2009; Rueda et al., 2009; Zhu et al., 2011; Liu et al., 2014; Zell et al., 2014a; French 65 66 et al., 2015; Hanna et al., 2016; Sinninghe Damsté et al., 2016; Warden et al., 2016) and open ocean 67 settings (Huguet et al., 2008; Fietz et al., 2012; Sparkes et al., 2015; Pan et al., 2016; Yamamoto et al., 68 2016; Jaeschke et al., 2017).

The global distribution of brGDGTs in soils and peatbogs has been related to both pH and air temperature (Weijers et al., 2007; Peterse et al., 2009a, 2012; Tyler et al., 2010; Naafs et al., 2017b). These environmental influences on brGDGT distributions have been expressed through the measurement of the degree of cyclisation of the brGDGTs, calculated via the so-called CBT index, and the degree of methylation, expressed as the MBT or MBT' indices. The global calibration of these indices yields significantly large errors of about 5.0 °C and 0.8 pH units (Weijers et al., 2007; Peterse et al., 75 2012), which can be reduced in some regional data sets (Tierney et al., 2010; Loomis et al., 2012; 76 Anderson et al., 2014; Yang et al., 2014b; Wang et al., 2016; Coffinet et al., 2017). The empirical 77 relationship between MBT and air temperature has been further validated by the passive incubation of 78 experimental peatland plots (Huguet et al., 2013b), and the study of soil samples in altitudinal or 79 geothermal gradients (Sinninghe Damsté et al., 2008; Peterse et al., 2009b; Ernst et al., 2013; Liu et al., 80 2013; Anderson et al., 2014; Deng et al., 2016; Coffinet et al., 2017). Similarly, the CBT vs. pH 81 relationship in soils was confirmed through the study of brGDGT distributions in long term (>45 years) 82 soil pH manipulation plots (Peterse et al., 2010).

Given the temperature and pH dependence of brGDGT indices in soils and peats, they are used as 83 84 proxies for these parameters in loess-palaeosol, peat and sedimentary records (e.g. Weijers et al., 2007; 85 Schouten et al., 2008; Rueda et al., 2009; Ballantyne et al., 2010; Bendle et al., 2010; Tyler et al., 2010; Fawcett et al., 2011; Peterse et al., 2011; Gao et al., 2012; Niemann et al., 2012; Zech et al., 2012; Ajioka 86 87 et al., 2014b; Sanchi et al., 2014; Cao et al., 2017; Wang et al., 2017a). However, brGDGTs are also 88 argued to be produced *in situ* within the water column or the sediments of aquatic settings (e.g. Tierney 89 and Russell, 2009; Tierney et al., 2010; Loomis et al., 2011; Buckles et al., 2014). This poses a major 90 challenge or may invalidate the application of the soil brGDGT proxies and their calibration equations 91 in sedimentary environments to reconstruct air temperatures and soil pH. However, in lakes such a 92 drawback is circumvented by the development of regional lacustrine sediment calibrations (Blaga et al., 93 2010; Tierney et al., 2010; Zink et al., 2010; Pearson et al., 2011; Peterse et al., 2011; Sun et al., 2011; 94 Buckles et al., 2014). In fact, there is also a significant correlation on a global scale of the CBT/MBT 95 proxies in lacustrine sediments with temperature and pH (Sun et al., 2011), but new proxy models are 96 still necessary in order to improve the accuracy and precision of using brGDGT distributions as climatic 97 proxies (Pearson et al., 2011). However, the identity and niche of the source organisms in aquatic 98 settings are still unknown.

In some regional studies, pH and temperature were not the primary variables that explained the spatial distribution of brGDGT indices. Their variability may have been influenced by precipitation, soil properties (texture, humidity), seasonality of production, and/or the use of air rather than soil temperature in the calibrations (Weijers et al., 2007; Peterse et al., 2009a; Rueda et al., 2009; Loomis et al., 2011; Peterse et al., 2011; Sun et al., 2011; Dirghangi et al., 2013; Anderson et al., 2014; Menges et
al., 2014; Wang et al., 2014; Dang et al., 2016; Davtian et al., 2016; Lei et al., 2016). Some studies have
argued that the original MBT/MBT' and CBT indices, if modified to account for the identification of
new brGDGT isomers, might explain some of the scatter in the global calibrations (De Jonge et al.,
2014b; Yang et al., 2015; Wang et al., 2016), but this may not hold true universally (Warden et al.,
2016). The choice of extraction technique, that recovers different pools of lipids, may also lead to a bias
in the observed distributions of brGDGTs (Huguet et al., 2010b; Weber et al., 2017).

110 The brGDGT palaeo proxies were initially calibrated against annual mean values of environmental 111 parameters. A number of studies have claimed that the proxy estimates are biased towards specific 112 seasons, usually summer mean values (Rueda et al., 2009; Peterse et al., 2011; Shanahan et al., 2013; 113 Wu et al., 2013; Deng et al., 2016; Wang et al., 2016). This hypothesis relies on the assumption that 114 there is a preferential period of bio-production of brGDGT, due to favourable environmental conditions, 115 that might further cause a seasonal bias in the temperature-related membrane lipids in the soils or peats 116 (Huguet et al., 2013b). However, in general, no apparent seasonal pattern in the distribution of either 117 core (C-brGDGTs) or intact polar lipid (I-brGDGTs, i.e., GDGTs bound to 1 or 2 polar head groups) 118 derived brGDGTs has been found in soils, exhibiting no obvious shift towards a particular season 119 (Weijers et al., 2011; Lei et al., 2016). The main argument to explain the lack of seasonality in soil 120 derived signals is the slow turnover time of brGDGTs in soils. This has been estimated to be of ca. 18 y 121 for branched C-GDGTs in arable soils in temperate climates (Weijers et al., 2010), and 8-41 y in peat for C-brGDGTs and 11–14 y for I-brGDGTs, suggesting that brGDGTs turnover is on timescales of 122 123 decades in terrestrial environments (Huguet et al., 2017). However, even much longer turnover times 124 (320–510 y) were observed in anoxic samples (Huguet et al., 2017).

Few studies have monitored the seasonal variability of brGDGT proxies, by means of the monthly sampling of SPM or settling particles in a trap in lakes (Blaga et al., 2011; Buckles et al., 2014; Loomis et al., 2014) or the ocean (Yamamoto et al., 2016). In general, while the fluxes of brGDGTs are seasonal and coupled to the local climatology, the estimated brGDGT temperatures did not match the monthly instrumental variability in local air or water temperatures. In some instances, the fractional abundances of brGDGTs in settling particles showed little seasonal variability over the sampling period (2-3 years) despite large seasonal changes in the fluxes and the actual air and water temperatures (Loomis et al.,
2014; Yamamoto et al., 2016).

133 In this study, we assessed further the occurrence of seasonality in the production and distribution of 134 C-brGDGTs by undertaking their monthly analysis for over a year in both soils and particulate matters from three different catchments. Two sites were located in mountain environments in the pre-Pyrenees 135 136 and the Pyrenees, and one site in southern Norway (Fig. 1a). The locations considered all have a marked seasonal temperature cycle, but they afforded different types of particulate matter and soils from 137 138 catchments of different size and nature, which we expected would maximize the possibility of detecting any seasonal bias in the production and composition of brGDGTs. Thus, the high mountain site in the 139 140 Pyrenees was chosen to obtain recently formed suspended particulate matter in the runoff collected in 141 ephemeral streams from the spring melting of the seasonal snow cover on a small catchment overlapping 142 a former glacier cirque. The Norwegian site, in contrast, was selected to obtain particulate matter 143 suspended in the waters of the longest and largest river in Norway that drains an area also with seasonal 144 snow cover and high precipitation levels. The third site considered provided settling particulate matter 145 from a trap in a lake from a mid-mountain catchment with a Mediterranean vegetation cover, which is 146 only occasionally snow covered a few days a year. As I-brGDGTs were expected to reflect contributions 147 from living biomass at the time of sampling (Weijers et al., 2011), we also analyzed the occurrence of 148 I-brGDGTs from samples in two of the river catchments (Ulldeter and Øsaker).

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150 2. Study sites and methods

151 **2.1. Sample collection**

152 **2.1.1. NE Iberian Peninsula**

153 One set of samples was collected in the Lake Montcortés catchment (42°19'50'' N, 0°59'41'' E, 1,027 154 m a.s.l.), located in the southern Catalan pre-Pyrenees (Fig. 1c). The catchment lies on a karstic terrain 155 composed of Triassic materials (Rosell, 1994). There is no meteorological station in Lake Montcortés. 156 Consequently, meteorological data were obtained from a nearby station located in the town of La Pobla 157 de Segur (585 m a.s.l.), which is about 10 km from the study area. To take into account the higher 158 elevation of Montcortés Lake, we estimated the rate of temperature change with altitude in the study 159 region as 1 °C per 100 m in elevation. This was calculated using air temperature data from three 160 meteorological stations near Montcortés in the towns of Sort, El Pont de Suert and La Pobla de Segur 161 from September 2013 to May 2015. The annual mean air temperature (MAT) was estimated to be 8.5 °C, ranging from -3.3 °C in January to 17.6 °C in July during the sampling period at the elevation of 162 Montcortés. For the period 1961-1990, the average accumulated annual precipitation in Pobla de Segur 163 164 was 669 mm, with February as the driest month (33 mm) and May as the wettest month (88 mm). However, during the sampling year (2014), the precipitation was more variable, with March as the driest 165 166 month (25 mm) and August as the wettest (155 mm) (Fig. 2a). The catchment is covered by 167 evergreen/deciduous oak forests, conifers, grasses and littoral vegetation (Rull and Vegas-Vilarrúbia, 2015). From September 2013 to November 2014, soil samples (Ms1, Ms2 and Ms3) were obtained 168 169 monthly from three different sites with different vegetation cover. For each soil type, three replicates 170 (at least 5 m apart from each other within a 10 m radius area) were collected from an area of 400 cm² and a depth of 5 cm, after removing leaf litter and clearing the surface from any visible traces of 171 172 vegetation. The soils were placed in a pre-combusted aluminium tray, wrapped with aluminium foil and 173 sealed in plastic bags. Settling particles were collected in Montcortés Lake using a PVC tube, capped at 174 one end, with a diameter of 8.5 cm and deployed at 20 m depth. Every month the trap was retrieved, and 175 its contents were filtered through Whatman glass fibre filters (GF/F, 0.7 µm pore diameter). Samples 176 were frozen on arrival at the laboratory until analysis.

177 A second series of measurements was undertaken in samples from Ulldeter, a glacial cirque in the 178 Catalan Pyrenees (Fig. 1c), and the headwaters of the Ter river. From June 2010 to May 2011, the MAT 179 (from a local meteorological station at 42°25' N, 2°15' E, 2364 m a.s.l.) in Ulldeter was 3.3 °C, ranging 180 from -2.8 °C in January to 12.7 °C in July. The accumulated annual precipitation was 1264 mm, with 181 October as the wettest month (228 mm) and February as the driest month (44 mm) (Fig. 3a). The area 182 was covered with snow from November to March. The temperature and precipitation data in Ulldeter 183 (Servei Meteorològic de Catalunya, 2011) during the studied periods are shown in Fig. 3. Soil samples 184 were collected in a flat terrain, 100 m northwards from the meteorological station. Three soil replicates

were taken each time, at least 5 m apart from each other within a 10 m radius area. The first 3 cm, 185 186 corresponding to plant litter, were removed and the 2 cm of topsoil was sampled from a surface area of 187 100 cm². Samples were stored in plastic bags and frozen at -20 °C on arrival at the laboratory until 188 analysis. The source of Ter River is close to the sampling site. In spring and summer 2011 two sediment 189 traps were installed for a month: one was in an ephemeral stream that fed the Ter River (Trap A, 2300 190 m a.s.l.) running down from a mountain slope, the other was deployed in a small meander in the Ter 191 River (Trap B, 2274 m a.s.l., attached to the stream bed). The trap contents were filtered through 192 Whatman glass fibre filters (GF/F, 0.7 µm pore diameter) and stored frozen at -20 °C until analysis.

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2.1.2. Southern Norway

195 Samples were collected near Øsaker, in a research farm located in the municipality of Sarpsborg situated 300 m away from Lake Vestvannet, which drains into the Glomma River (Fig. 1b). The 196 197 meteorological station is located at the sampling site (59°19' N, 11°02' E, 45 m a.s.l.). During the 198 sampling year (December 2010 – November 2011), MAT was 6.0 °C, ranging from -10.9 °C in December to 17.1 °C in July (Fig. 4a). The soil temperature at 10 cm depth showed a similar variability 199 200 to air temperature in Øsaker, but the average annual soil temperature (7.7 $^{\circ}$ C) was 1.7 $^{\circ}$ C higher than 201 local MAT. The accumulated annual precipitation was 1033 mm, with September (236 mm) as the wettest month and December (20 mm) as the driest month (NIBIO Agro MetBase, 202 203 http://lmt.bioforsk.no). Each month soils were sampled in a forested area, 70 m away from the 204 meteorological station. The top, corresponding to plant litter, was removed using a pick and a shovel. 205 The first 2 cm of topsoil were sampled from an area of 100 cm². The sampling site for the water 206 collection is located 6 km southeast of the meteorological station and 16 km upstream from the Glomma 207 River mouth to the Skagerrak. Water samples (10 L of water) were collected from the Glomma River 208 0.5 m under the surface at Sandesund (Sarpsborg; 59°16' N, 11°05' E, 2.5 m a.s.l.) each month from 209 February 2011 to January 2012. Water samples were sequentially filtered through Whatman glass fibre 210 filters (GF/F, 0.7 µm pore diameter) for GDGT analysis. Samples were frozen on arrival at the 211 laboratory.

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213 **2.2. Sample preparation**

214 Montcortés soils (2-8 g, dry weight) were extracted with a mixture of 30 mL of 215 dichloromethane:methanol (DCM:MeOH, 3:1, v/v, ×3) using a CEM-MARS microwave extractor 216 (Kornilova and Rosell-Melé, 2003). After cooling to room temperature, the total extract was evaporated 217 to near dryness and then fractionated over an aminopropyl silica column with 12 mL DCM:isopropanol (2:1, v/v; F1) and 15 mL diethyl ether: acetic acid (96:4, v/v; F2). Fraction F1 was then eluted through a 218 219 silica column using hexane (6 mL, n-alkanes), DCM (4 mL, alkenone) and DCM:MeOH (95:5, v/v, 6 220 mL, GDGTs). The fraction containing GDGTs was dried with a vacuum centrifugal concentrator, re-221 dissolved with hexane: isopropanol (99:1, v/v), and filtered before instrumental analysis.

Filters from sediment traps of the Montcortés lake were extracted in a sonication bath with 20 mL of DCM:MeOH (3:1,v/v, x3), and centrifuged. Supernatants from the same sample were combined and dried by rotary evaporation. The extract was fractionated with 0.6 g silica (230-400 mesh; deactivated with 1% H₂O) using hexane (5 mL), DCM (4 mL), DCM:MeOH (95:5, v/v; 6 mL; GDGT fraction) and MeOH (5 mL). The GDGT fraction was fractionated further through an aminopropyl-silica (0.8 g; 4 0-60 μ m particle size) column using DCM:isopropanol (2:1, v/v; 8 mL). The extract was dried and filtered through a 0.45 μ m PTFE filter prior to instrumental analysis

229 Soil and river water samples from Ulldeter and Øsaker were prepared according to the procedure 230 described by Huguet et al. (2010b). The samples were extracted in an ultrasonic bath using MeOH (\times 3), 231 DCM:MeOH (1:1, v/v; ×3) and finally DCM (×3). After each sonication, samples were centrifuged and 232 all the supernatants were combined, and taken to dryness using a vacuum centrifugal concentrator. The 233 lipid extracts were fractionated with 0.5 g of activated silica and eluted with hexane:DCM (9:1, v/v; 3 234 mL; F1), hexane:DCM (1:1, v/v; 3 mL; F2) and DCM:MeOH (1:1, v/v; 3 mL; F3; GDGTs). The last 235 fraction containing both core GDGTs (C-GDGT) and intact polar GDGTs (I-GDGT) was divided into two parts. One part was filtered through a 0.45 µm PTFE filter and stored until instrumental analysis. 236 237 The other half was hydrolyzed to remove the polar head groups and to allow quantification of both core 238 and intact polar lipids (Huguet et al., 2010b). Hydrolysis was performed with 2 mL of 5% HCl in MeOH (v/v) at 70 °C for 4 h. GDGTs were recovered by a liquid-liquid extraction using DCM and deionized water (1:1, v/v). The recovered DCM extracts were combined and rinsed six times with deionized water to remove any traces of HCl. The extracts were dried and filtered through a 0.45 µm PTFE filter prior to instrumental analysis.

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244 2.3. Instrumental analysis

245 GDGTs in Montcortés samples were analyzed by injecting 5-20 µL aliquots of the fractionated 246 extracts in a high performance liquid chromatograph (HPLC; Agilent 1200 RRLC) Time-of-Flight 247 (TOF) mass spectrometer (MicroTOF-Q, Bruker Daltonics) with an atmospheric pressure chemical 248 ionization (APCI) interface set in a positive mode. GDGTs in Ulldeter and Øsaker samples were 249 automatically injected in a Dionex P680 HPLC coupled to a Thermo Finnigan TSQ Quantum Discovery 250 Max quadrupole mass spectrometer. Following the methods (Escala et al., 2007; Huguet et al., 2013c), 251 the extracts were eluted through a Tracer Excel CN column (Teknokroma; 20 cm length, 0.4 cm diameter 252 and 3 µm particle size) equipped with a pre-column filter and a guard column. The mobile phase was 253 eluted with hexane:n-propanol (98.5:1.5, v/v) at a flow of 0.6 mL/min and the parameters of the APCI 254 interface were set as follows to generate positive ion spectra: corona discharge 3 µA, vaporizer 255 temperature 400 °C, sheath gas pressure 49 mTorr, auxiliary gas (N_2) pressure 5 mTorr and capillary temperature 200 °C. GDGTs were detected in selected ion monitoring (SIM) mode at (Schouten et al., 256 257 2007): m/z 1050, 1048, 1046, 1036, 1034, 1032, 1022, 1020, 1018 and 1208 (abbreviated as GR, which was used as the internal standard). Prior to extraction, GR standard that is a synthetic tetraether lipid 258 259 with a structure typical of neutral archaeal membrane lipids (Réthore et al. 2007) was added to the sample. The reproducibility of the quantification of GDGTs using GR as a standard is higher than 90% 260 (Escala et al., 2009). 261

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263 **2.4. Ancillary measurements**

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2.4.1. Water content and LOI

265 The total amount of organic carbon (TOC) in soil samples was estimated using the loss on ignition (LOI) method (Schumacher, 2002). Aluminium trays were combusted at 450 °C for 8 h. About 1 g of 266 267 newly collected soil sample (in triplicate) was weighed (WW) in a combusted aluminium tray and dried (18 h) at 105 °C in the oven. The dried samples were weighed (DW105) and heated at 550 °C in the 268 269 muffle furnace for 4 h and left to cool down to room temperature before weighing them (DW₅₅₀). The 270 LOI value was calculated using the equation as follows (Heiri et al., 2001): 271 Water content (%) = $(WW - DW_{105})/WW \times 100\%$ 272 LOI (%) = $(DW_{105} - DW_{550})/DW_{105} \times 100\%$ 273 The TOC was calculated according to generally accepted TOC/LOI ratio of 0.58 (De Vos et al., 2005): 274 275 TOC (%) = $0.58 \times LOI \times 100\%$ 276 2.4.2. pH analysis 277

The freeze-dried soils were homogenized and sieved (2 mm mesh size) to remove roots and small 278 279 stones. The pH of the soil was potentiometrically measured in a supernatant suspension of a 1 to 2.5 280 soil: water mixture (Thomas, 2000). About 10 g of sieved soil was placed into a wide-mouth polythene 281 container and mixed with deionized water (25 mL). The container was capped and shaken 1 h with a reciprocating shaking machine. A pH meter (GLP22, Crison Instrument) with a glass-calomel 282 combination electrode was calibrated with a buffer solution at pH 4.0 and 7.0. The samples were shaken 283 for about 5 seconds and the sensing electrode was immersed inside to record the value when it was 284 285 stabilized (1σ , 0.1 pH unit).

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287 **3. Results**

288 **3.1. Sites characteristics**

The three sites showed clear differences and a marked interannual variability, in air temperature and rainfall during the sampling period (Figs. 2a, 3a, 4a). Specifically, there was a marked difference in the temperature range between the coldest and hottest month, with the amplitude of the range larger in Øsaker (28.0 °C) and Montcortés (20.9 °C) than in Ulldeter (15.5 °C). The mean annual rainfall was the highest in Ulldeter (1264 mm), followed by Øsaker (1034 mm) and Montcortés (793 mm) (Fig. 2a, 3a, 4a). The difference between the wettest and driest month was 184 mm, 215 mm and 136 mm, respectively.

296 The soil TOC varied substantially between the three sites. The lowest TOC contents were found in 297 the two sites in the Pyrenees, ranging from 6.4 to 17.7 % (mean 11.6 ± 1.0 %) in Montcortés and from 7.6 to 34.3 % (mean $20.3\pm7.7\%$) in Ulldeter (Table 2,3). In contrast, the TOC values were more than 298 299 twice as high, and fluctuated from 37.8 to 53.6 % (mean 49.3±4.1%) in Øsaker (Table 1). These results 300 were in line with the spatial distribution of TOC in Europe, with much higher values of organic carbon 301 in Scandinavia than in Southern European topsoils (Jones et al., 2004; De Brogniez et al., 2015). During 302 a year, the TOC content of soils in Montcortés and Øsaker fluctuated in a range of 6.8% and 15.8%, 303 respectively. Ulldeter soils showed larger ranges, with the highest TOC in February/March and the 304 lowest in May (Supplementary material Table S3). TOC in soils was negatively correlated with air 305 temperature ($R^2=0.71$, p<0.01), and positively correlated ($R^2=0.96$, p<0.01) with monthly accumulated 306 amount of rainfall (<150 mm/month).

The soils from Montcortés, which had the lowest TOC values, were also the least acidic, with pH values that ranged from 7.0 to 7.9 (mean 7.3 ± 0.5), typical of calcareous forest soils. Conversely, the soils with the highest TOC had the lowest pH values, from 4.8 to 6.8 (mean 5.5 ± 0.6) in Ulldeter, and in Øsaker with values between 3.9 and 4.5 (mean 4.1 ± 0.1), typical of cool-temperate humid deciduous or mixed deciduous and coniferous forest soils (Supplementary material Tables S2-4) (Sumner, 1999). The dominant soil types in three sites were Halciustoll (Ulldeter), Torriorthent (Montcortés, Ms3), and 313 Podzolicsoil (Øsaker) according to the soil classification system published by Sumner (1999). It is

314 apparent that the seasonal variability of soil pH values in these three sites was very small.

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316 **3.2. brGDGT concentration and distribution in soils**

317 Concentrations of C-brGDGTs (normalized to TOC) varied along the sampling period within the 318 same orders of magnitude in the three sites, albeit higher average values occurred in Øsaker. In the 319 Montcortés catchment they ranged from 5.5 to $21.8 \,\mu g/g_{TOC}$ (mean $11.4 \pm 4.9 \,\mu g/g_{TOC}$), from 2.3 to 20.1 320 $\mu g/g_{TOC}$ (mean 10.9±4.7 $\mu g/g_{TOC}$) in Ulldeter soils, and from 3.3 to 73.2 $\mu g/g_{TOC}$ (mean 27.5±24.8 $\mu g/g$ 321 TOC) in Øsaker soils (Fig. 2b, 3b, 4b). Previous studies reported that soils showed concentrations of C-322 brGDGTs from 1.2 to 10.9 μ g/g TOC in soils from the Lake Challa catchment in east Africa (Buckles et 323 al., 2014), from 1.3 to 17.5 μ g/g _{TOC} in the Iberian Peninsula (Menges et al., 2014) and from 0 to 12 μ g/g 324 TOC in Chinese soils (Yang et al., 2014a). In addition, all the C-brGDGT concentrations are in the range 325 of mid-latitude soils without seasonal variation (Weijers et al., 2011).

326 The I-brGDGT (not analyzed in Montcortés) concentrations were also within the same order of 327 magnitude in the two sites, but higher in Øsaker on average. In Ulldeter, the I-brGDGT abundance 328 showed similar values (4.8-24.4 μ g/g_{TOC}, mean 9.7±6.9 μ g/g_{TOC}) to C-brGDGT from December to May, 329 and lower values (<2.8 μ g/g TOC, mean 1.3±0.9 μ g/g TOC) than C-brGDGT (>5.0 μ g/g TOC) from June to 330 November (Fig. 3b). In Øsaker the soil I-brGDGT concentrations were similar to C-brGDGT 331 concentrations, ranging between 0.6 and 56.2 μ g/g TOC with a mean value of 20.2±20.1 μ g/g TOC, showing higher concentrations from March to June. Our data seem to contrast with those of previous studies 332 which reported that I-brGDGTs comprised only a smaller part of the total pool of branched GDGTs (e.g. 333 Weijers et al., 2009; Peterse et al., 2010), as both pools of lipids have similar range of values in our 334 335 samples. Our data do not seem to support the notion that production of brGDGTs from the living biomass 336 is warm-season biased (e.g. Deng et al., 2016). In contrast, the highest concentration of I-brGDGTs in 337 Ulldeter soils in December indicated a signature towards cold temperatures. In comparison with Ulldeter 338 soils and soils from other sites where C-brGDGTs are dominant (Weijers et al., 2009, 2011; Huguet et al., 2010b; Tierney et al., 2012), the concentrations of I-brGDGTs in Øsaker soils varied within a larger
range of values.

341 The distributions of C-brGDGTs and I-brGDGTs were nearly homogenous over time in the soils from 342 the same site, but were dissimilar among the different sampling sites (Fig. S2). The dominant brGDGTs (Supplementary material Fig. S1) in Montcortés soils for sites Ms1, Ms2 and Ms3 were IIa, IIIa and IIb 343 344 (>70%), followed by Ia, Ib. BrGDGT-IIa, Ia and IIIa were generally the most abundant compounds (55-345 84%) in Ulldeter soils, followed by IIb, Ib. BrGDGT-Ia and IIa were the most abundant ones (>90%) in 346 Øsaker soils, followed by IIIa. The average distribution pattern in Montcortés and Ulldeter soils is 347 similar to those in Tagus river catchment (Zell et al., 2014b) and in Øsaker soils is similar to that of soils 348 from the Svalbard archipelago in the Arctic area reported by Weijers et al. (2007). The 349 distribution/relative abundance of each compound in the same soil did not change much over the sampling year (Supplementary material Tables S2-4), as it will be discussed when analyzing the 350 351 variability of the brGDGT ratios.

352

353 **3.3. brGDGT fluxes and distributions in particulate matter**

354 In Montcortés Lake, the aim was to monitor the C-brGDGT variations every month in vertical settling 355 flux of particulate matter. The lowest fluxes occurred from January to June, and the highest fluxes from July to December (Fig. 2a), varying from 0.59 ($\mu g/m^2/day$) in February/March to about 8.45 ($\mu g/m^2/day$) 356 357 in September/December. The observed trend in the annual cycle of the fluxes in Montcortés resembles those described for the temperate lakes Lower King Pond in Vermont (USA) (Loomis et al., 2014) and 358 lake Lucerne in Switzerland (Blaga et al., 2011), and two tropical lakes (Challa and Huguangyan Maar) 359 360 (Buckles et al., 2014; Hu et al., 2016). The magnitudes of the ranges in the C-brGDGT fluxes were also 361 within the same order of magnitude in the 5 different lakes, being 0.1-20.0 μ g/m²/day in the Lower King Pond, between 0.84-3.0 ng/m²/day (42 m) and 0.87-11.7 ng/m²/day (72 m) in lake Lucerne; 0-12.0 362 $\mu g/m^2/day$ in the Challa lake, and 3.2-24.2 $\mu g/m^2/day$ in the Huguangyan Maar lake. 363 In Ulldeter, the aim was to collect recently eroded soil particulate matter, carried by runoff from 364

365 snowmelt and after episodes of rainfall. Given that the ground was frozen and covered with snow for 5

366 months, the sampling was only undertaken during spring and summer, when ephemeral streams appeared 367 in the mountainside below the soil sampling site. The accumulated abundance of C- and I-brGDGTs in 368 spring (0.60 and 0.34 μ g/g dry weight) was higher than in summer (0.26 and 0.15 μ g/g dry weight), 369 which might be due to co-effects of snowmelt and precipitation in spring (Fig. 3a).

370 In Øsaker, given the proximity of the Glomma River to the soil sampling site, the particulate matter 371 derived from the runoff was obtained from the filtration of water samples in this river. The samples containing C-brGDGTs during the study period varied in concentration from 1.6 ng/L in February to 372 373 54.3 ng/L in April. The abundance of I-brGDGTs was much lower than C-brGDGTs in Glomma River, ranging between 0.2-3.0 ng/L in 9 out of the 12 samples, similar to the distribution of these brGDGTs 374 375 (I-brGDGTs and C-brGDGTs) in Columbia River (French et al., 2015). The highest I-brGDGT 376 concentrations were found in March and December (above 21.9 ng/L; Fig. 5d). Both maxima in concentrations of C- and I-brGDGTs were offset by 1-2 months with the mean river discharge, that 377 peaked from May to July in Glomma River at Solbergfoss (50 km north from Sandesund, where the 378 379 samples were collected) during the period 1901–2007 (Global Runoff Data Centre, 2011) (Fig. 5d).

The Montcortés trap, during the sampling period, had a similar distribution of brGDGTs, on average, to the soils studied (Supplementary material Fig. S2, Table 2). BrGDGT-IIa and IIIa were the most abundant homologues (59-67 %) followed by brGDGT-Ia, IIb and Ib, and traces of IIb, IIc, and Ic, throughout the sampling period.

In the Ulldeter traps, the brGDGTs distributions were also relatively similar to those in soils, where brGDGT-IIa was generally the most abundant compound in the SPM, followed by Ia, and IIIa, and at lower abundance IIb and Ib (Supplementary material Fig. S2). However, the fractional abundance of brGDGT-IIIa increased by >5% on average in the traps compared with soils, while in the intact polar fraction, brGDGT-IIIa increased by more than 10% in trap B, deployed in Ter River.

In the Glomma River, the differences in the composition of brGDGTs in soils and SPM were significant. Thus, in both core and intact polar fraction, the IIa was the most abundant compound instead of the most abundant Ia in soils. In addition, the fractional abundance of IIIa increased by 5-10 times compared with that of soils, while the proportion of IIa and IIb only slightly increased. In contrast, the proportion of Ia was reduced to a half, and brGDGT-Ib, Ic and IIc were still present in smaller amountin both soils and the SPM (Fig. S2).

The distributions of C- and I-brGDGTs in Glomma SPM are similar to that of I-brGDGT in Yenisei river gulf and mouth SPM that showed decreased Ia and increased IIIa (De Jonge et al., 2015).

397

398 **3.4. GDGT proxy ratios**

399 The ratios measured derived from brGDGTs were the following:

400 MBT = [Ia+Ib+Ic]/[Ia+Ib+Ic+IIa+IIb+IIc+IIIa+IIIb+IIIc] (Weijers et al., 2007) (eq.1)

401 MBT' = [Ia+Ib+Ic]/[Ia+Ib+Ic+IIa+IIb+IIc+IIIa] (Peterse et al. 2012) (eq.2)

402 $CBT = -Log[(Ib+IIb)/(Ia+IIa)] = 3.33-0.38 \times pH (n=114; R^2=0.7) (Weijers et al., 2007) (eq.3)$

The C-brGDGT CBT and MBT soil values from the three sites were within the range of values in global soil compilations, although they are substantially different among each other. Thus, the values from our three sites span almost the whole range of reported CBT values so far and half the range of MBT values (Figs. 6a and 6b).

407 In Montcortés, the CBT values of C-brGDGTs (C-CBT values) in soils ranged from 0.19 to 0.50, and 408 in settling particulate from 0.34 to 0.52, a similar range to the values in the soils in the catchment (Fig. 409 2). In Ulldeter soils, the C-CBT values varied over a wider range of values, from 0.52 to 1.54, which 410 encompasses the values in the trap A of 0.84/0.83 (spring/summer), and in the trap B of 1.03/0.99 411 (spring/summer) (Supplementary material Table S3). The CBT values of I-brGDGTs (I-CBT values) 412 ranged from 0.56 to 1.53 in Ulldeter soils and were highly correlated to C-CBT values (R²=0.85, 413 p<0.01). In the traps from Ulldeter the spring/summer I-CBT values were 0.77/0.78 (spring/summer) in 414 the trap A and 0.99/0.98 (spring/summer) in the trap B, which are within the range of the soil values. In 415 Øsaker soils, C-CBT values varied from 1.40 to 1.69 (with one extreme value of 0.77 in March), and 416 from 0.83 to 1.68 in Glomma River SPM. The range of I-CBT values was from 1.16 to 1.66 in Øsaker 417 soil, and from 0.93 to 1.35 in Glomma River SPM.

The MBT values ranged from 0.20 to 0.35 in Montcortés soils, from 0.25-0.42 in Ulldeter soils and
from 0.49-0.62 in Øsaker soils. MBT from C-brGDGTs ranged from 0.21 to 0.26 in Montcortés traps

420 that showed low values from April to July and higher values from August to March. MBT from brGDGT 421 core lipids were 0.32/0.25 and 0.27/0.27 in Ulldeter Trap A and Trap B, in spring and summer, 422 respectively, while MBT from I-brGDGTs were 0.31/0.32 and 0.24/0.24. MBT from brGDGT core 423 lipids and intact polar lipids were within the range of 0.29-0.37 in Glomma River SPM over a year 424 except for C-MBT in November and I-MBT in February that showed higher values around 0.60. 425

426 **4. Discussion**

427 **4.1. Controls on brGDGT abundance**

428 No significant correlations were observed between total brGDGT concentrations in the soils from the 429 three sites and MAT, MAP (mean annual precipitation) or pH, indicating that no single factor was 430 responsible for the brGDGT abundances in our samples. However, specific regional conditions may 431 sometimes play a significant role in brGDGT occurrences. For example, under the same local regime of 432 precipitation and temperature, the brGDGT concentrations in soils (normalized to TOC) varied by an 433 order of magnitude, even in a small catchment such as the ones in Montcortés lake. Such variability 434 suggests that precipitation and temperature do not directly influence brGDGT production in soils. 435 However, the brGDGT abundance seemed to be somewhat related to soil water content in Montcortés 436 soils ($R^2=0.43$, p<0.05, Fig. 7a). This correlation improved when we expanded the number of soil 437 locations investigated in the catchment and used the average annual values of brGDGT concentration 438 (when available) ($R^2=0.87$, p<0.05, Fig. 7b). This finding is consistent with that observed in many 439 regions that the concentrations of brGDGTs correlate positively (strongly or moderately) with soil water 440 content, and that humid soils may provide a more favourable environment for brGDGT production than 441 drier soils, under the same climatic regime (e.g. Dirghangi et al., 2013; Yang et al., 2014a; Ding et al., 2015; Wang et al., 2017b). Other studies also showed that soil water content had a clear impact on the 442 absolute concentrations of all brGDGTs, with higher abundance in wetter soils (e.g. Dang et al., 2016). 443 444 This is also assumed to explain the result that the higher brGDGT abundance occurs mostly in Øsaker

soils, which are more humid than the other two sites even though we did not analyze the soil watercontent in Ulldeter and Øsaker soils.

447 The brGDGT abundance/flux of particulate matter in the river/lake showed distinctive seasonal 448 variations due to different transport drivers (Fig. 5). In Montcortés, the drivers of the brGDGT flux 449 during the annual cycle were likely to be wind and precipitation. The former mode of transport, carrying 450 soil particles to the lake, is operational throughout the year although likely to be more intense in summer 451 (Supplementary material Table S2, Fig. 5b). In contrast, rainfall runoff will drive sediment input to the 452 lake most significantly when there is a rainstorm (Corella et al., 2016). The brGDGT flux peak in 453 December might be due to water column mixing (Fig.5b) (Trapote et al., 2018), as reported in other 454 temperate and tropical lakes (e.g. Buckles et al., 2014; Hu et al., 2016). During the partial lake turnover, 455 'fossil' brGDGTs from a pool suspended and preserved in the lower water column might be resuspended 456 and trapped, increasing the brGDGT flux in the trap. Conversely, in Ulldeter, spring snowmelt runoff 457 was also likely to carry soil matter to the local streams, as well as precipitation in summer. Arguably 458 these would be the periods with the highest loads of brGDGT in soil-eroded matter and settling 459 particulate matter in the Ter River. In the Glomma River, the brGDGT abundance reached a maximum 460 one month (April) before the highest discharge (May) that was driven by snowmelt and precipitation 461 (Fig. 4a, 5c). The brGDGT abundance in the river seemed not to be directly related to mean monthly 462 river discharge (Fig. 5d). Our data suggest that the observed variation in SPM in the Glomma River 463 should result from a mixing of soil-derived and *in situ* produced brGDGTs, leading to the highest flux of brGDGT supply from the riverine in *situ* produced brGDGTs. 464

465

466 **4.2. Regional environmental controls on brGDGT proxies**

The mean values of the CBT-brGDGT index in our study and their correlation with soil pH values fits well with the trends previously observed from the compilation of global soil data (Fig. 6, e.g. Weijers et al., 2007). Thus, CBT is mainly controlled by pH on large spatial scales (regional or global scale; $R^2=0.72$, p<0.01), but the trends were not so significant at small scales (Fig. 6a). In contrast, the correlation between MBT values and pH is not so significant on a global scale ($R^2=0.46$; Fig. 6b). 472 However, if a subset of data is considered, using data from previously reported Iberian soils (Menges et 473 al., 2014) and those from this study, a significant correlation of MBT with pH is also apparent ($R^2=0.77$, 474 p<0.01) (Fig. 6b). A significant positive correlation between MBT' and soil water content has been 475 reported previously, based on C6-methylated brGDGTs (e.g. Dang et al., 2016; Naafs et al., 2017b). 476 Instead, Naafs et al. (2017b) have established new global calibrations for temperature based on the 477 finding that temperature is the primary control on C5-methyl brGDGTs in soils. Chromatographic separation and quantification of the C5- and C6-methylate isomers have improved the error statistics of 478 479 temperature and pH calibrations based on brGDGTs in soils through the removal of the interference of 480 the C6-methyl compounds from C5-methyl brGDGT fractional abundances (De Jonge et al., 2014b). 481 Weber et al. (2015) analyzed sediment and watershed soils from a Swiss lake and documented the 482 presence of a novel "mixed" hexa-methylated brGDGT that was only detected in the lake and not in 483 soils of the surrounding watershed. This finding provided further evidence for *in situ* production of 484 brGDGTs in lakes, increasing the possibility that separation of the 5- and 6-methyl isomers could be 485 more useful. However, in our study, we did not separate C5 and C6-methylated brGDGTs, and the MBT' 486 was generally higher in more humid soils (e.g. Ms1>Ms3).

487 Changes in soil pH and soil water content could explain a bias in the reconstructed MAT. The precise 488 influence of different environmental factors on the estimation of MAT values is likely to be difficult to 489 establish and will depend on local conditions in different sites or regions, or analytical improvements. 490 For instance, in an extensive set of soils from humid, semi-humid to semi-arid and arid regions in China, 491 the MBT/CBT derived temperatures matched the measured MATs better in humid and no-alkaline 492 regions (Zheng et al., 2016). In our study, MAT was particularly underestimated in the estimates derived 493 from the brGDGT indices in the more alkaline soils from Montcortés (Ms3) where pH values were 494 higher. In contrast, values from the acidic Øsaker soils were overestimated (Fig. 8). It seems that soil 495 water content and pH might constrain the suitability of brGDGT-proxies for continental air temperature 496 reconstruction in some instances or locations, which affects the errors of any calibration. However, the 497 overall global relationship is clearly strong. Thus, after the combination of our results with the data set 498 published by Peterse et al. (2012) in Fig. S3, it is clear that the relationship between estimated and 499 instrumental MAT is very close to a 1:1 line. As argued, any efforts to reduce the large errors in the

MBT/CBT-estimated MAT (Weijers et al., 2007; Peterse et al., 2012) should probably focus in regional
calibrations efforts (e.g. Anderson et al., 2014; Yang et al., 2014b). Moreover, chromatographic
separation of the 5-methyl and the 6-methyl isomers may improve the accuracy of the MAT estimates
(De Jonge et al., 2014; Naafs et al., 2017).

504

505 **4.3. Seasonality of the MAT estimates**

Our data corroborate previous claims on the absence of seasonal patterns in the abundance and 506 distribution of C-brGDGTs and I-brGDGTs in soils (e.g. Weijers et al., 2011). For instance, the 507 reconstructed temperatures from C-brGDGTs (C-MAT_{es} = $0.81-5.67 \times CBT+31 \times MBT'$; Peterse et al. 508 509 2012) remained rather constant throughout the year in the Montcortés soils, and thus did not follow any 510 climate seasonal variations (Fig. 8a). In the three sites of the catchment, C-MAT_{est} values in soils ranged 511 between 7.4-11.3 °C, 7.1-9.9 °C and 1.0-5.1 °C (Fig. 8a). In fact, some of the differences in the mean 512 values of C-brGDGTs among the three Montcortés soils sites were larger than the annual variability 513 within each site (Fig. 8a). Thus, the average C-MAT_{est} reconstructed from samples Ms1 and Ms2 were 514 9.1±1.1 °C and 8.1±1.0 °C, respectively, equivalent to the instrumental MAT of 8.5 °C. While C-MAT_{est} 515 reconstructed from Ms3 was 3.1±1.0 °C, which was significantly colder than the instrumental MAT. 516 This bias could be due to increased pH and aridity in Ms3, whose soils were located in a dry alkaline area. The lack of seasonality in the soil C-MAT_{est} was also mirrored by the data from the settling 517 particulate matter in Montcortés lake. The C-MAT_{est} ranged between 4.1 and 6.8 °C, with an average of 518 519 5.1±0.7 °C, lying within the range of the three soil end members aforementioned. In fact, the annual variability in C-MAT_{est} of the settling material in the lake is smaller than in soils. This phenomenon 520 521 might be indicative that C-MAT_{est} from materials in the sediment trap originated from the catchment 522 soils, and that the C-MAT_{est} in the lake sediments is mainly the weighted average from different soil inputs in the catchment. 523

524 The absence of a meaningful seasonal trend C-MAT_{est} in soils from the Ulldeter site is also apparent, 525 with values ranging between 0.1 and 7.8 °C (mean 3.6 ± 2.2 °C) (Fig. 8b). At Ulldeter, the values of C-526 MAT_{est} from the sediment Trap A (Fig. 8b), located at a higher elevation, also lacked a seasonal trend,

with values higher in spring (6.0 °C) than summer (3.7 °C). These values were in fact closer to MAT 527 528 (3.3 °C) than to real spring (2.3 °C) or summer temperature (10.2 °C). The C-MAT_{est} for the sediments 529 deposited in Trap B in the Ter River during spring and summer seasons were 3.4 °C and 3.5 °C, respectively (Fig. 8b), less variable than soil C-MAT_{est} and also close to Ulldeter instrumental MAT (3.3 530 °C). The average value of the estimates from I-brGDGTs (i.e. I-MAT_{est}), which represents the inputs 531 532 from living biomass at the sampling time, was 2.8 °C, 0.5 °C lower than instrumental MAT and 0.8 °C lower than soil C-MATest. The similarity of the C-MATest data from both traps confirms the expectation 533 534 that collected sediment resuspended in stream waters from the same source, which was eroded from original soils and the C-MAT_{est} signal is representative of the overall standing stock of core lipids in 535 536 soils.

Interestingly, the I-MAT_{est} and C-MAT_{est} values also showed a different variability pattern during the 537 538 sampling period in the soils and in the traps (Fig. 8b), the variability being larger for I-MAT_{est} than C-539 MAT_{est} . Such differences would endorse the proposal that these signals are derived from two different 540 lipid pools, one associated to relic material (C-brGDGTs) and the other dominated by inputs from living 541 biomass (I-brGDGTs). This was further supported by the few data that we obtained on I-brGDGTs from Trap A, where I-MAT_{est} was 6.0 °C in June and increased to 10.5 °C in August, resembling the 542 543 instrumental air temperatures rather than the signal from C-brGDGTs. The June I-MAT_{est} was 0.8 °C 544 lower than the instrumental June temperature, while the August I-MAT_{est} was only 0.6 °C above the 545 instrumental August temperature (11.1 °C). These data suggested that the distributions of I-brGDGTs in 546 river particles may contain the new seasonal production signal from recently eroded soils or aquatic environments, which were not available from the measurement of C-brGDGTs. Thus, a seasonal bias in 547 548 the MAT_{est} would be obtained from the measurement of I-brGDGTs but not from C-brGDGTs in river 549 particles, that would instead yield estimates closer to the actual MAT values. I-brGDGTs were under 550 limit of quantification in Trap B from the Ter River samples.

In Øsaker soils, the C-MAT_{est} ranged between 8.3 and 14.7 °C with an average of 10.7 ± 1.5 °C, which is 4.7 °C above the instrumental MAT in Øsaker. Similarly, the I-MAT_{est} ranged between 7.3 and 11.1 °C with an average of 9.5±1.2 °C (Fig. 8c). Consequently, there was no significant difference between C- or I-brGDGT based MAT estimates in soils. As in the other two sites, the amplitude of the seasonal

cycle in the MAT in Øsaker was not reflected in the variability of the soil brGDGT MAT signatures. 555 556 These have similar range of variability annually in the three sites, of ± 5.1 °C in Montcortés (average), 557 ±6.4 °C in Øsaker, and ±7.7 °C in Ulldeter. In comparison, the instrumental MAT variability is ±20.9 558 °C in Montcortés, ±28.0 °C in Øsaker, and ±15.5 °C in Ulldeter (Supplementary material Table S2-4). 559 Nevertheless, in Øsaker, the annual C-MAT_{est} is also substantially warmer than the instrumental MAT. 560 while in Ulldeter and Montcortés the C-MAT_{est} was close to the instrumental value. This 'warm' signal 561 in Øsaker soils might be due to the decreased fractional abundance of C6-methyl brGDGTs as the 'cold' underestimated MAT was observed in arid soils with a large amount of C6-methyl brGDGTs (De Jonge 562 563 et al., 2014b). The Glomma River is the longest and largest river in Norway, with a basin area of 42,000 564 km², which drains 13% of Norway's area. Its course has a change in elevation of 690 m. Consequently, 565 the sediment particles sampled near its mouth, obviously, will have a much wider range of sources than 566 the soils taken in the Øsaker station. In that sense, there was no reason to expect that the brGDGTs 567 signatures in the SPM should have the same patterns to surrounding soils as described so far for 568 Montcortés and Ulldeter. In fact, the variability of C- and I-MATest in Glomma River SPM was larger 569 than in the Øsaker soils (Fig. 8c). However, we observed the same lack of seasonal pattern in C-MAT_{est} 570 already described in the settling particles of Montcortés Lake or the runoff from Ulldeter, which was 571 also apparent in the I-brGDGT signatures. Moreover, the average C- and I-MAT_{est} in the Glomma River 572 particulate material was 5.0/5.2 °C, just 1 °C below the instrumental MAT at Øsaker, and lower than 573 soil C-MAT_{est}. Such differences were rather small and might be attributed to the sources of the particles further upstream than Øsaker, in locations at higher elevations and with a continental climate (Rueda et 574 575 al., 2009). These data further supported the notion that the standing stock of core lipids in soils was 576 much higher than any new annual brGDGT production over a seasonal cycle, and that the slow turnover time, on timescales of decades, of brGDGTs occurred in terrestrial environments (e.g. Weijers et al., 577 578 2010; Huguet et al., 2017). This may provide an explanation as to why the bulk of the soil brGDGT 579 signature is in fact largely representative of average annual air temperatures, independently of the annual 580 amplitude of the local temperature seasonal cycle. Consequently, the bulk of the eroded soil signal 581 represented by the brGDGTs, being carried to river courses and eventually to a sedimentary

environment, would not be seasonally biased and could be best used to estimate MAT in the regionsinvestigated.

584

585 **4.4. Particulate matter brGDGT sources**

In the areas investigated in our study, we propose that the main sources of brGDGTs in the particulate matter are the soils from the lake or stream catchment. This interpretation is based on the observations that, i) the brGDGT distributions in soils and the particulate matter have more points of similarity than otherwise, and, ii) the lack of seasonality in the proxy signals in the three settings, both in the soils and particulate matter signals. The latter argument assumes that the turnover time of brGDGTs in particulate matter would be on seasonal time-scales rather than decades or centuries as reported for soils and peatlands (e.g. Weijers et al., 2010; Huguet et al., 2017).

593 However, this conclusion would appear to contradict the well-established notion in the literature that, 594 in aquatic settings, brGDGTs are mainly derived from *in situ* sources, rather than allochthonous ones. 595 This notion has been sustained, mainly, on the description of different brGDGT distributions in 596 sediments in relation to those in soils on adjacent lands (e.g. Sinninghe Damsté et al., 2009, 2016; 597 Bechtel et al., 2010; Loomis et al., 2011, 2014; Sun et al., 2011; Tierney et al., 2012; Schoon et al., 2013; 598 Ajioka et al., 2014a; Buckles et al., 2014; Naeher et al., 2014; Colcord et al., 2015; Weber et al., 2015; 599 Warden et al., 2016; Zink et al., 2016). In addition, some brGDGTs have only been identified in 600 lacustrine sediments that had a distinct carbon isotopic signature from brGDGTs in soils (Weber et al., 601 2015). Fietz et al. (2012) also argued that the widespread significant correlation between concentrations 602 of brGDGTs and crenarchaeol, a common archaeal biomarker in aquatic settings (e.g. Schouten et al., 603 2013), could be interpreted as evidence of *in situ* production of brGDGTs in lacustrine and marine 604 settings. Moreover, the occurrence of the intact form of brGDGTs in living cells (I-brGDGT) across the 605 continuum from soil to marine environments can also be construed as evidence for the widespread in 606 situ production of brGDGTs in a wide range of environments (e.g. French et al., 2015).

However, some studies have also noted that in some instances the distributions of brGDGTs in a lake
catchment soils and sediments are similar (e.g. Niemann et al., 2012; Hanna et al., 2016). Peterse et al.

609 (2015) observed no in situ production of brGDGTs, but did observe production of archaeal (i.e. 610 isoprenoidal) GDGTs, in 160-day incubations of soils in fresh and ocean water. In the Colville, Congo 611 and Danube river contributions from in situ derived brGDGTs were not deemed significant in relation 612 to the signal derived from local soil inputs within the drainage basin and ultimately exported to the sea 613 (Hanna et al., 2016; Freymond et al., 2017; Hemingway et al., 2017). In the Baltic Sea, significant in 614 situ production was also deemed unlikely given the correlation between brGDGTs and terrigenous 615 indicators (Kaiser and Arz, 2016). These findings and our study would show that, even though in situ 616 production may be widespread, it is not necessarily always the dominant source of brGDGTs in all 617 aquatic settings. Clearly, this poses an added difficulty in the choice of the most appropriate calibration 618 in sedimentary palaeorecords (i.e. soil vs. sediments), which can be compounded if the relative inputs 619 from allochthonous vs autochthonous sources change as well through time.

620 Regarding the lack of seasonality in the brGDGT proxies in the settling particulate matter, as already 621 noted, this is not the first study to make such an observation. Loomis et al. (2014), in a small and 622 relatively shallow kettle lake in northern Vermont argued that the lack of seasonality was caused by a 623 dominance of inputs into the trap (at 6.5 m depth) from the lake subsurface rather than from surface 624 production of brGDGTs, or from surrounding soils. In our study in the Montcortés trap, we do not have 625 any evidence that significant *in situ* production of brGDGTs is occurring in the Montcortés lake. We 626 propose that the simplest explanation is that the main source of brGDGTs is the lake catchment, from 627 its soils. The non-seasonal signals in the soils are transferred into the settling particles in the trap, which 628 have a distribution of brGDGTs that appear to be a weighted mean of the soil signals (Fig. 8a). It is not 629 surprising that, even in a small catchment as in Lake Montocortés, the soils are spatially chemically 630 heterogenous. By their very nature many soils are heterogenous even at small spatial scales (e.g. Jackson 631 and Caldwell, 1993). How the different soils from a catchment may contribute to the settling particle 632 signals may depend on, i) the facility with which each site erodes, ii) its location in the catchment so the 633 soil particles via runoff reach the lake, and, iii) the concentrations of brGDGTs in each soil location, 634 which is correlated to soil water content in the Montcortés catchment (Fig. 7). The relative importance 635 of different areas of the catchment as a source of brGDGTs may change over time, through change in 636 land uses and/or climatic/weather events (Corella et al., 2011, 2016).

Moreover, other potential sources that may deserve further investigation are the aeolian inputs into the lake. Dust is composed of soil and sediment particles that contains brGDGTs (e.g. Fietz et al., 2013), and can contribute significantly to the exported sedimentary signal to the deep sea (e.g. Yamamoto et al., 2016; Jaeschke et al., 2017). In continental settings, eolian inputs of brGDGTs have not been taken into consideration to explain sedimentary or loess/paleosols sequences except in a very few instances (e.g. Zech et al., 2012). Further research should focus on the degree to which atmospheric redeposition, by the eolian transport of GDGTs, might overprint local signals in a particular site.

644

645 **5. Conclusions**

No seasonal trends were observed in the distribution of brGDGTs in soils from environments with a large amplitude in the seasonal cycle of temperature and precipitation. Results confirmed previous findings from mid-latitude soils, on the lack of a detectable short-term seasonal signal on the background soil brGDGT distributions. Moreover, we show that the brGDGT signatures in particulate materials derived from the same catchment also lacked a seasonal pattern.

Conversely, the brGDGT production in different types of soils from different catchments varied during an annual cycle. The temporal variability of brGDGTs is not coupled to the annual cycle of temperature and precipitation, nor is transferred to the brGDGT proxy indices, i.e. CBT-MBT('). In fact, the dependence of these proxies with annual mean pH and air temperature is robust, particularly when analyzed at large spatial scales (i.e. continental). Moreover, this climatic signal is transferred to the sediments derived from runoff in the catchment as the particulate matter in the three environments had very similar MAT_{est} values in soils.

The interpretation of the reconstruction of any proxy requires an understanding of the spatial and temporal scales on which the reconstructed parameters are representative of. In our study, we provide further evidence that the brGDGT signatures in depositional environments are representative of the soils in the hydrological catchment. In addition, we demonstrate that the sedimentary signal is unlikely to contain a seasonal bias when derived from soil sources. The accuracy of brGDGT-derived temperature estimates based on MBT/CBT proxy in soils will depend on soil properties rather than seasonal signals. Furthermore, soils are likely to contain a brGDGT signature which is representative of average environmental conditions in the catchment over decades if not longer time scales, depending on the soil type and the age of the carbon stocks. Consequently, brGDGT proxy reconstructions based on soil and peat sources of brGDGTs should be considered to infer variability in environmental parameters over the same timescales (i.e. decades or longer). On shorter timescales (i.e. annual), sediment downcore variability is likely to be related to changes derived from *in situ* production and sediment sources.

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680	Figures
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Figure 1. a) General location of the study areas; b) Øsaker, southern Norway; c) Montcortés and
Ulldeter, Catalan Eastern Pyrenees. Maps obtained from the software Ocean Data View (Schlitzer,
2010). The circles indicate the locations of sampling sites.

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Figure 2. a) Monthly accumulated precipitation (bars) and instrumental monthly mean air temperature (lines) in Pobla de Segur (10 km from Lake Montcortés); b) Average concentrations of C-brGDGT (brGDGT core lipids) in the Lake Montcortés catchment and; c) Average brGDGT-derived CBT and MBT indices in Montcortés soils from September 2013 to November 2014. The error bars show the range for the values from three different soils in the catchment.

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Figure 3. a) Monthly accumulated precipitation (bars) and instrumental monthly mean air temperature
(lines) in Ulldeter; b) Concentrations of core (C-brGDGT) and intact polar (I-brGDGT) branched
GDGTs in the Ulldeter soils and; c) C-brGDGT and I-brGDGT derived CBT and MBT indices in
Ulldeter soils from June 2010 to May 2011.

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Figure 4. a) Monthly accumulated precipitation (bars) and instrumental monthly mean air temperature
(lines) in Øsaker; b) Concentrations of core (C-brGDGT) and intact polar (I-brGDGT) branched GDGTs
in Øsaker soils; and c) C-and I-brGDGT derived CBT and MBT indices in Øsaker soils from December
2010 to November 2011.

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Figure 5. a) Temperature and precipitation in Montcortés (Pobla de Segur); b) Flux of core branched
GDGTs in the sediment trap in Lake Montcortés from October 2013 to November 2014. c) Temperature
and precipitation in Øsaker; d) Concentration of brGDGTs in suspended particulate matter (SPM) from
the Glomma River. Concentration of C-brGDGTs (hollow diamonds) and I-brGDGTs (crosses) are

compared to mean discharge (m³/s, discontinous line) of Glomma River at Solbergfoss for the period
1901-2007 (Global Runoff Data Centre, 2011).

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Figure 6. Linear regression plots of global data on (a) measured soil pH vs. CBT, and (b) measured soil
pH vs. MBT indx data. Data plotted in black and blue are from Weijers et al. (2007) and Menges et al.
(2014), respectively, data plotted in other colours are from this study.

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Figure 7. The relationship between soil water content and core brGDGT concentration in Montcortés catchment: a) Values from the monthly sampling over the study period (September 2013-November 2014) from sites Ms1, Ms2 and Ms3, characterized by grassland, oak forest and upland dry grassland/scrubland, respectively. b) Values of 6 soil sites around the Montcortés lake, where Ms1, Ms2, Ms3 represent the average values during the study period, whereas sites Ms4, Ms5, Ms6 were only sampled once in January 2016.

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Figure 8. Meteorological data and GDGT based temperature estimates for soils from Montcortés (a), 720 721 Ulldeter (b) and Øsaker (c). Monthly instrumental mean temperature is given as MMAT (black 722 continuous line with crosses) and average instrumental values during the study period as MAT (black 723 dashed lines). The results for the GDGT based estimates (MBT'/CBT) based on core GDGTs (C-724 MATest, solid diamonds) and intact GDGTs (I-MATest, hollow diamonds) are shown for the monthly 725 sampled soils. The results for the sediment traps (in Montcortés lake and Ulldeter streams) and water 726 samples (in Glomma River) are shown with blue diamonds (C-MATest), and hollow blue diamonds (I-727 MATest).

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Figure















