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1 **Impact of grass silage quality on greenhouse gas emissions from dairy and beef**
2 **production**

3 Running title: Silage quality and GHG emissions

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14

15 **Abstract**

16 High quality grass silages may represent a mitigation option by reducing enteric methane
17 production and by increasing productivity, thus reducing greenhouse gas emissions per kg of
18 product (emission intensity). Two previous studies found considerable effects of three different
19 silage qualities cut at different maturity stages (very early (H1), early (H2) and normal (H3))
20 offered ad libitum with various levels of concentrate supplementation, on animal performances
21 of growing/finishing bulls and dairy cows in early lactation, indicating that emission intensities
22 may also vary. Based on results from these previous studies, the aim of this study was to
23 estimate emission intensities for milk and beef carcasses for the included combinations of
24 silage qualities and concentrate levels, by using the farm-scale model HolosNor. The emissions
25 intensities were lowest for the H1 silage, and highest for the H3 silage, independent of
26 concentrate levels for both milk and beef. Thus, increasing concentrate levels did not
27 compensate for lower grass silage quality. Improvements in silage quality from H3 silage to
28 H2 is realistic and has the potential to reduce emission intensities with approximately 10%
29 while keeping the milk yield per cow constant and reducing the use of concentrates
30 considerably. For beef production the potential is even larger, with a reduction in emission
31 intensity of approximately 17%. We conclude that improving grass silage quality may be a
32 mitigation option that will also reduce the dependence on concentrates.

33

34 **Keywords:** greenhouse gas emissions, grass silage quality, dairy production, beef production,
35 farm scale models; emission intensity

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41 1 INTRODUCTION

42 Grasslands are an important land use in Europe, and permanent grasslands cover about 8% of
43 the land area and 35% of the agricultural area, with large geographical variations (Smit et al.,
44 2008). Grasslands are especially important (i.e., proportion of permanent and temporary
45 grasslands of total agricultural land >50%) in parts of Western Europe (the Netherlands,
46 Luxembourg, Ireland, Scotland and Wales), the mountainous areas of Central Europe (Austria,
47 Montenegro, Slovenia and Switzerland), the Mediterranean area (Greece, Macedonia, Bosnia
48 and Montenegro), the Caucasus (Georgia and Azerbaijan) and Northern Europe (Iceland and
49 Norway) (Smit et al., 2008). In the latter, climate and topography restrict areas suitable for
50 agriculture, crop production especially (FAO, 2012). In the Nordic countries, agricultural land
51 area is only 9% of the total land area, of which 38% is used for grain production (country-
52 specific values ranging from 0% on Iceland to 56% in Denmark). Meadows and pastures covers
53 48% of the agricultural area, ranging from 30% in Finland to 98% in Iceland. Even so, the use
54 of these grass resources has decreased during the last decades, and the use of concentrates for
55 ruminants is substantial and increasing (Åby et al., 2014). However, expected human
56 population growth, climate change that may lead to more challenging production conditions
57 resulting in reduced yields for important food and feed crops in tropical and temperate regions,
58 and increased competition with other land use, such as biofuels and urban expansion, may pose
59 risks to global food security and limit the availability of grains for animal feeds (IPCC, 2014;
60 FAO, 2006; Nordic Statistics, 2016). Thus, production systems for ruminants which is mainly
61 based on grass and less dependent on concentrates may be of importance for maintaining a high
62 degree of self-sufficiency in many regions.

63 Grass silage is the main winter feed for both dairy and beef production in Norway (Randby et
64 al., 2010). There is a potential to improve grass silage quality, here defined as grass silage
65 nutritive value, which is mainly obtained by cutting the grass at an earlier maturity stage
66 (Harrison et al., 1994). This may increase net energy intake from grass, thus obtaining
67 increased productivity, for example higher milk yield per cow and growth rates of fattening
68 bulls, while decreasing the use of concentrates.

69 The effects of varying grass silage qualities and levels of concentrate supplementation on feed
70 intake and performance of growing/finishing dairy bulls and dairy cows in early lactation was
71 investigated by Randby et al. (2010, 2012). These authors used three grass silage qualities
72 (denoted H1, H2 and H3, where the maturity stage at harvest corresponded to very early, early
73 and normal maturity stage, respectively) offered *ad libitum* with different levels of concentrate
74 supplementation. Maturity stage at harvest was found to correlate with grass silage quality, and
75 a variation from 6.75 MJ net energy/kg DM for the very early maturity stage to 5.52 for the
76 normal maturity stage was found. Animal performances varied considerably between
77 treatments, demonstrating the effects of improved grass silage quality. For example, dairy bulls
78 were finished before 15 months of age on H1 silage. Average daily milk yield during lactation
79 week 1-16 was highest using H1 silage with 8 kg concentrates (32.8 kg energy corrected milk
80 (ECM)), however relatively high yield was obtained solely on this silage (H1; 23.4 kg ECM).
81 H3 silage led to a maximum of 30.1 kg ECM, when supplemented with the optimal level of
82 concentrate, 12 kg. Corresponding grass silage DM intake for the treatments were 16.7, 16.9
83 and 11.9 kg, respectively.

84 Greenhouse gas (GHG) emissions from ruminants are important to consider due to its effects
85 on climate change (FAO, 2006), and mitigation strategies to reduce the environmental impact
86 are of high interest. Several nutritional strategies have been suggested to reduce methane
87 emissions (Beauchemin et al., 2008). Hristov et al. (2013a) proposed improving forage quality
88 as one of the most efficient ways of decreasing CH₄ emissions, through increased production

89 efficiency thereby diluting the maintenance energy requirement and reducing the number of
90 animals needed to produce the same amount of product (Boadi et al., 2004; Hristov et al.,
91 2013b). Improving forage quality may reduce enteric methane production due to lower fibre
92 and/or higher soluble carbohydrates content. Improved forage quality may also increase
93 voluntary intake, reducing the retention time in the rumen and reducing the proportion of
94 dietary energy converted to methane (Eckard et al., 2010). On the other hand, cutting grass at
95 an early maturity stage reduces the dry matter (DM) yield, which may increase the need for
96 grassland areas (e.g., Kuoppala et al., 2008) and thereby increase the use of fertilizers, leading
97 to higher N₂O emissions, and increased use of fossil fuels. Thus, when looking into mitigation
98 options, it is crucial to use a whole farm approach, to ensure that emissions do not increase
99 elsewhere in the production chain as pointed out by Eckard et al. (2010).
100 The considerable effects of improved grass silage quality on animal performances found by
101 Randby et al. (2010) and Randby et al. (2012), gives reason to believe that emission intensities
102 (GHG emissions per product; milk and finished young bull carcass) may also vary. Thus, the
103 objective of this study was to investigate if improved grass silage quality reduces emission
104 intensities by using the results from these studies for the included combinations of grass silage
105 qualities and concentrate levels as inputs in the farm scale model HolosNor (Bonesmo et al.,
106 2013) to calculate emission intensities for both milk and beef carcass.

107

108 **2 MATERIALS AND METHODS**

109 Emission intensities for milk and young bull carcass were calculated based on the results from
110 the feeding experiments of Randby et al. (2012) and Randby et al. (2012), and a short summary
111 of the studies are given below (section 2.1 and 2.2). In addition, several assumptions were made
112 in order to do the calculations in the HolosNor-model, described in section 2.3.

113

114 **2.1 Grass silages used**

115 The two studies were performed simultaneously in the same barn with the same feeds. Five
116 leys used for silage preparation were sown with the same seed mixture, consisting of 50%
117 timothy, 35% meadow fescue and 15% red clover. All swards were fertilised 26.-27. April with
118 69 kg N, 13 kg P and 33 kg K/ha. Each ley was divided into three parts, consisting of
119 approximately 50%, 30% and 20% of the area, for harvesting at the three different maturity
120 stages/harvesting dates, respectively: 30 May to 1 June (H1), 6-8 June (H2) and 14-16 June
121 (H3), corresponding to 6.75, 6.26 and 5.52 MJ net energy lactation/kg grass silage DM,
122 respectively. Compared to Norwegian practice, H1, H2 and H3 corresponds to harvesting at
123 very early, early and normal maturity stages for timothy, respectively. The silage DM yield per
124 ha were 3,350, 5,210 and 6,250 kg for H1, H2 and H3, respectively.

125

126 **2.2 Animal performances**

127 The three grass silage qualities were fed to Norwegian Red dairy cows during early lactation,
128 2-3 weeks before expected calving date to week 16 in lactation (Randby et al., 2012) and to
129 growing finishing Norwegian Red bulls from age 7 months until slaughter (Randby et al.,
130 2010).

131

132

133 2.2.1 Dairy cows

134 Cows were held in a 3 x 3 factorial arrangement with the three grass silage qualities
135 supplemented with three levels of concentrates (4, 8, 12 kg/day, denoted C4, C8 and C12). In
136 addition, H1 was offered as a sole feed and H3 with 16 kg of concentrates/day, giving in total
137 11 diets studied. The dietary treatments are denoted H1C0, H1C4, H1C12,.....H3C16. From
138 the observed average daily milk yield in week 1-16 by Randby et al. (2012), 305-days yields
139 (Table 1) was approximated by comparing the average daily milk yield in week 1-16 from the
140 standard lactation curve as used by the Norwegian dairy cooperative TINE SA:

$$141 \quad Kg \text{ FPCM}_{DIM} = 0.993 + (0.00312 * 305yield - (0.0984 * DIM) + (LN(DIM * 3.726)))$$

142 Where FPCM is fat- and protein-corrected daily milk yield, DIM is days in milk (1,2....305),
143 305yield is 305-day milk yield in kg, LN is the natural logarithm

144 The milk composition (fat and protein %) for the included grass silage qualities and concentrate
145 levels was as found by Randby et al. (2012) (Table 1).

146

147 2.2.2 Growing/finishing bulls

148 All bull calves were given hay and grass silage *ad libitum*, 4-8 l acidified milk during the first
149 3 months, and up to 1.5 kg of concentrates per day. Average daily growth rate pre experiment
150 was 1,036 g. From age 7 months, bulls were divided into six groups and given the three grass
151 silage qualities *ad libitum* as a sole feed, or with a daily supplementation of 2 kg concentrates,
152 increasing to 3 kg at 385 kg live weight (LW) and 4 kg at 500 kg LW. The bulls were
153 slaughtered at approximately 575 kg LW. Age at slaughter, LW and total concentrate
154 consumption for the included combinations of silage qualities and concentrate levels are given
155 in Table 2.

156

157 2.3 Calculation of greenhouse gas emissions using HolosNor

158 Emissions intensities, kg CO₂-equivalents per kg FPCM or finished young bull carcass, for the
159 three grass silage qualities and concentrate levels were calculated using the farm scale model
160 HolosNor (Bonesmo et al., 2013). This model estimates GHG emissions from dual-purpose
161 milk and beef production systems and considers the direct and indirect emissions of CH₄, N₂O
162 and CO₂ from direct and indirect sources. The direct emissions result from on farm livestock
163 production activities such as enteric fermentation and production of roughage, while the
164 indirect emissions are from inputs used on farm such as fuel and fertilisers, and nitrate leaching
165 and volatilization. In addition, soil C changes are estimated (Bonesmo et al., 2013). Enteric
166 methane emissions in kg are calculated on the basis of an IPCC Tier 2 approach. Gross energy
167 (GE, MJ) is multiplied by the methane conversion factor (Y_m, proportion of methane of total
168 GE intake) divided by the energy content of methane (55.64 MJ/kg) (Bonesmo et al., 2013).
169 Gross energy intake is calculated from the net energy requirements (IPCC, 2006; NRC, 2000;
170 NRC, 2001) for all animal groups and taking into account the energy density of the diet.
171 HolosNor adjusts Y_m to account for the digestibility of the dietary dry matter (DM) where Y_m
172 = 0.1150-0.0008×DE% (Bonesmo et al., 2013). For example, Y_m values of 5.8, 6.1 and 6.6%,
173 were calculated for the H1, H2 and H3 silage, all supplemented with 4 kg concentrates,
174 respectively. Farm characteristics used as inputs in HolosNor were from various sources,
175 described below. Animal performances were based on Randby et al. (2010) and Randby et al.
176 (2012) as described above. Weather and soil data used were from Ringsaker municipality

177 (Skjelvåg et al., 2012), an important dairy region in the Eastern parts of Norway. In order to
178 compare the different treatments at the same level of milk production, a target milk production
179 on farm was determined based on the current average herd size in Norway (26 cows), and an
180 average milk yield of 7,100 kg ECM per cow and year (Statistics Norway, 2016). The number
181 of dairy cows needed to fulfil this target for all treatments were calculated based on the 305-
182 days yields in Table 1. HolosNor requires input on the time spent on pasture for dairy cows
183 and heifers, however as the data from the experiments did not include grazing, time spent on
184 pasture was set to zero. The ley areas needed for all treatments (Table 3) was estimated based
185 on the total grass yield per ha, and the total grass silage requirement for all animal groups. Total
186 grass yields were calculated assuming two cuts for the H3 quality, and three for H1 and H2
187 (Bakken et al., 2009). Based on the results from a large field study (Bakken et al., 2009) it was
188 assumed that yields for the second and third cut for H1 was 90 and 74% of the first cut, while
189 it was 56 and 48% for H2. For H3, yield of the second cut was assumed to be 78% of the first
190 cut. Thus, the total silage yield for H1, H2 and H3 was 8,860, 10,620 and 11,120 kg DM/ha,
191 respectively. Grass silage requirements was calculated by HolosNor on the basis of net energy
192 requirements (IPCC, 2006; NRC, 2000; NRC 2001), as functions of herd specific data such as
193 animal performances and the number of animals in all groups, after subtracting the energy
194 intake from concentrates (Table 3). Concentrate use was calculated separately and was an input
195 into the model. Total concentrate consumption for dairy cows was calculated as the total feed
196 requirements as a function of milk yield (Volden, 2013) corrected for the observed grass silage
197 intake by Randby et al. (2012). For growing fattening bulls, concentrate consumption from age
198 7 months to slaughter was given by Randby et al. (2010). In addition, it was assumed a
199 concentrate consumption of 228 kg DM per bull before 7 months of age, based on the feed
200 recommendations of Berg & Matre (2007). Concentrate net energy value was 6.56 MJ per kg
201 DM. The emission intensity for purchased concentrate was calculated from the amount of
202 grains (barley produced off-farm) and imported soybean meal needed to supply the energy and
203 crude protein used. Emission intensities for barley and soy bean meal was, 0.62 and 0.93 kg
204 CO₂-eq/kg DM, respectively. Land use change was not included in these figures (Bonesmo et
205 al., 2013). Estimates of soil C change were based on the Introductory Carbon Balance Model
206 (ICBM) (Andrén et al., 2004). The ICBM model estimates the change in young and old soil C
207 from total C inputs (sum of C in plant residues and manure), a humification coefficient, two
208 decay constants and the relative effect of soil moisture and temperature. Fertiliser use for all
209 silage qualities and cuts was 69 kg N per ha (Randby et al., 2010). A dairy cow replacement
210 rate of 30% (TINE, 2013) was used and the number of finished young bulls was 0.57 per cow
211 and year (calculated as a function of average values for slaughter age of cow, age at first
212 calving, calving interval and calf losses) (TINE, 2013). Average live weight of dairy cows was
213 539 kg and barn electricity consumption per cow was 1,720 KWh (Bonesmo et al., 2013). To
214 allocate emissions between meat and milk, a physical allocation method which reflect the
215 underlying use of feed by the animals to produce milk and meat, was used (Bonesmo et al.,
216 2013).

217 To calculate the effect of varying grass silage quality on total greenhouse gas emissions from
218 the cattle population (dairy and suckler), the current Norwegian production levels of 1,500
219 million liters of milk and 80,000 tons of beef (Åby et al., 2014) and the emissions intensities
220 from HolosNor was used. In order to limit the number of combinations of various grass silage
221 qualities and concentrate levels for milk and beef, two contrasting ones were chosen: H1C8
222 (highest yield) vs. H3C4 (low yield constrained by high dietary fiber concentration), combined
223 with the two levels of concentrate use in beef production. In addition H1C0 was included, to
224 investigate the effects of a completely grass silage-based system. Beef originating from the
225 dairy population was calculated from the number of dairy cows needed to meet the production
226 level of milk as a function of the 305-day yield (Table 1) and an annual beef production of 250

227 kg carcass per cow (Åby et al., 2016). Annual beef production per cow was calculated based
228 on average lifetime of cows, age at first calving, calving interval, calf loss and carcass weights
229 for young bulls, heifers and cows from slaughter statistics. A calve sex ratio of 1:1 was used.
230 One heifer calve was assumed kept as replacement. The discrepancy between the total
231 production level for beef and the beef production from the dairy population was assumed to be
232 from suckler cows. The emission intensity of suckler beef was assumed to be 25.5 kg CO₂-
233 equivalents per kg beef carcass, and annual beef production per cow was 277 kg (Åby et al.,
234 2016).

235

236 **3 RESULTS**

237 **3.1 Emission intensities for milk**

238 The lowest emission intensity was found for H1C8, but with very minor differences with H1C4
239 (+0.001 kg CO₂-equivalents per kg FPCM), while the difference with H3C4, the treatment with
240 the highest emission intensity, was 0.252 kg CO₂-equivalents (Figure 1). Moderate amounts of
241 concentrate supplements, up to 8 kg per day with H1 and H2 and 12 kg per day with H3,
242 reduced the emission intensity within grass silage quality. By contrast, the highest concentrate
243 level within all silage qualities, 12 kg for H1 and H2 and 16 kg for H3, increased the emission
244 intensity, but was still less than the lowest concentrate level (0, 4 and 4 kg, respectively) (Figure
245 1). Even so, emissions intensities were lowest for the H1 silage, and highest for the H3 silage,
246 independent of concentrate levels. For example, the emission intensity for H1 silage with 4 kg
247 of concentrates was lower than the H2 silage with 8 kg of concentrates and H3 silage with 12
248 kg concentrates. The combination H1 silage with no concentrates also had a lower emission
249 intensity than all concentrate levels within the H2 and H3 silages.

250 The effects of improved grass silage quality was larger when going from the H3 quality to H2
251 than from H2 to H1. For example, there was a 0.14 kg CO₂-equivalents/ kg FPCM reduction
252 from H3C4 to H2C4, while the reduction from H2C4 to H1C4 was 0.11 kg CO₂-equivalents
253 (Figure 1).

254 The most important emission sources per kg FPCM (Figure 2) were CH₄ from enteric
255 fermentation, CH₄ and N₂O from manure and N₂O from soils, included indirect emissions from
256 leaching and volatilization. H3C4 resulted in higher emissions from enteric fermentation and
257 manure, compared to H1C8 (Figure 2). There were small differences for soil N₂O and CO₂
258 from feed production and energy use between the two treatments, while soil sequestration was
259 higher for H3C4 (Figure 2).

260

261 **3.2 Emission intensities for finished young bull carcasses**

262 The highest emission intensity resulted from H3 without concentrate supplementation, while a
263 35% lower emissions intensity from H1 with concentrates was the lowest (Figure 3). For H1
264 and H2, the effect of concentrate supplementation was small, only a 0.25 and 3% reduction in
265 the emission intensity, respectively, but for but for H3 reduction was 14%. The effect of
266 improved roughage quality was largest from H3 to H2 without concentrates, with a reduction
267 of 4.2 kg CO₂-equivalents per kg finished young bull carcass. In comparison, the difference
268 between H2 and H1 without concentrates was 1.4 CO₂-equivalents per kg finished young bull
269 carcass. The emission intensity for H1 without concentrates was 9% lower than H2 with
270 concentrates.

271

272 **3.3 Total emissions**

273 The number of dairy and suckler cows needed in order to meet the domestic production targets
274 for milk and beef for the included grass silage qualities and concentrate levels differed
275 considerably (Figure 4). H1C0 had the lowest need for suckler, while H1C8 had an additional
276 need of 103,396 suckler cows in order to meet the beef production target. The lowest annual
277 total GHG emission was obtained from the combination of H1C0 in the dairy production with
278 H1 with concentrate supplementation for bulls, while the highest was from the H3C4 in dairy
279 production and H3 without concentrates for bulls, a difference of 788,772 tons CO₂-equivalents
280 (Figure 5).

281

282 **4 DISCUSSION**

283 **4.1 Grass silage quality, productivity, fertilizer use, area availability and profitability**

284 The superior grass silage quality (H1) gave the highest productivity (Randby et al., 2012;
285 Randby et al., 2010) and resulted in the lowest emission intensities in both dairy and beef
286 production independent of concentrate levels, as argued by Hristov et al. (2013a). Increasing
287 the concentrate level could not compensate for lower grass silage quality. For example, a milk
288 yield per cow of approximately 7,000 kg was obtained on H3 with 12 kg concentrates, H2 with
289 8 kg concentrates and H1 with 4 kg concentrates, while the emissions intensities were 0.943,
290 0.853 and 0.758 kg CO₂-eq., respectively. Similarly, a yield of approximately 6,300 kg per cow
291 was obtained with H3 with 8 kg concentrates or H2 with 4 kg concentrates. The emissions
292 intensities was 0.975 and 0.873, respectively. A realistic improvement from the average
293 Norwegian grass silage quality, which corresponds to the H3 silage, to H2 could thus reduce
294 the emission intensity by approximately 10%. At the same time, this will reduce the concentrate
295 use and increase the grass silage consumption (Table 3). Similarly, for young bull carcass, the
296 emission intensity may be reduced by approximately 17% when going from the H3 silage with
297 concentrates to H2 without concentrates. This indicates that improving grass silage quality may
298 be a potential mitigation option, while at the same time giving the opportunity to reduce the
299 use and dependence on concentrates, without reducing animal performance.

300 To the authors' knowledge, no other studies have estimated emission intensities using farm
301 scale models based on results from feeding experiments. However, studies using life cycle
302 assessments (LCA) have demonstrated diminished emission intensities with increasing animal
303 productivity both in dairy (e.g., Casey & Holden, 2005; Gerber et al., 2011) and beef production
304 (e.g., Capper, 2011; Wiedemann et al., 2015). Beauchemin et al. (2011) investigated the
305 mitigation potential in improved forage quality for a breeding stock of beef cattle during the
306 winter in a simulation study using a farm model LCA. Thus, in contrast to our study, the effect
307 of forage quality on emissions from fattening animals were not considered. The authors
308 assumed in their calculations that an earlier harvest date decreased grass yield (10% reduction),
309 increased DM digestibility and decreased Y_m, similar to our approach.

310 The lower grass yields of earlier harvested grass silage may be a challenge from a practical
311 point of view, as the farmer is dependent on a sufficient amount of grass silage for the long
312 indoor feeding season, which is approximately 8 months in Norway. Even so, there is a large
313 potential to increase yields through improved grassland management and agronomical
314 practices, and the grassland yield potential ranges from about twice the current yield in the

315 central and southwestern parts of Norway to 3.5 times in northern Norway (Bakken et al., 2014;
316 Steinshamn et al., 2016).

317 In this study, it was assumed that the same grass silage quality was obtained for all subsequent
318 cuts. This is of course a simplification and the grass silage quality of the subsequent cuts may
319 vary according to factors such as weather conditions, cutting regime etc. (Bakken et al., 2009).
320 The results are valid, however, to demonstrate the mitigation potential of improved grass silage
321 quality.

322 For simplicity, a fertilizer application of 69 kg N per ha for all cuts and grass silage qualities
323 was used in our calculations, even if N application is usually higher for the first cut (Bakken et
324 al., 2009). This gave a total N application of 207 kg N per ha for H1 and H2 (three annual cuts),
325 and 138 kg N for H3 (two annual cuts). This is lower than the recommendations of Yara
326 (2018). The average annual fertilizer application on grasslands in Norway is 177 kg N per ha,
327 but is higher (208 kg N per ha) for dairy farms, according to Bye et al (2016). In the studies of
328 Bakken et al. (2009), high grass silage qualities for the first cut corresponding to H1 were
329 obtained on a lower fertilizer level of 120 kg N per ha, however the total yield was 10-15%
330 lower compared to a fertilization level of 240 kg per ha. The assumptions on N application is
331 obviously important as it determines the N₂O emissions from soils. The N application assumed
332 in the present study and the resulting N₂O emissions for the H3 silage may be underestimated
333 compared to the average values given by Bye et al. (2016). Likewise, based on the higher
334 recommendations of Yara (2018), the N₂O emissions for the H1 and H2 silages may also be
335 underestimated.

336 As pointed out by Hristov et al. (2013a), profitability is the determining factor for the possible
337 adoption of any mitigation option. Bonesmo & Randby (2011) found that using the very early
338 harvested silage (H1) for fattening bulls only gave a marginal higher profit than the H2 silage.
339 Flaten et al. (2014) compared the profitability of differing harvesting regimes in dairy farming
340 and concluded that no harvesting regime is superior under all conditions, but that this depends
341 on the availability of land and other fixed farm resources such as milk quota and housing
342 capacity. High quality silages (H1) were only more profitable when there were no restrictions
343 in land availability and other fixed farm resources. Thus, from an economical viewpoint,
344 advocating the use of a very early harvested grass silage as a mitigation option may not be
345 preferable under current external production conditions.

346

347 4.2 Emission intensities

348 Reductions in emission intensity for milk with increased grass silage quality and concentrate
349 levels were mainly related to higher milk yield and thus fewer cows needed to meet the
350 production target (Table 1 and 3), thereby reducing methane emissions from enteric
351 fermentation and manure. The opposite effect (lower milk yield and a higher number of cows
352 to meet production target) explained the increase in emission intensity for the highest
353 concentrate levels within each grass silage quality. The differences between H1C8 and H3C4,
354 the treatments with the lowest and highest emissions intensities, respectively, were mainly
355 explained by differences in emissions of methane from enteric fermentation and animal
356 manure, while there were small differences for soil N₂O, and CO₂ from feed production and
357 energy use (Figure 2). Uptake of carbon in the soil was lower for H1C8 than for H3C4 (Figure
358 2). This was due to a smaller grass area (Table 3), a lower C residue yield (2,751 vs. 3,520 kg
359 per ha), less C from manure (2,890 vs. 3,004 kg per ha) and thus lower total C inputs to soil
360 (5,640 vs. 6,524 per ha). For the other emission sources, only small differences between the
361 two treatments was observed.

362 The variation in emission intensity of finished young bull carcasses (Figure 3) was closely
363 related to slaughter age, as a function of varying growth rates (Table 2). For H3, slaughter age
364 was reduced from 543 to 454 days with concentrate supplements, while it was only reduced
365 from 450 to 427 days for H1. Norwegian red bulls fed the H1 and H2 silages, were likely close
366 to their genetic potential for growth, which may explain the small effects on emission
367 intensities (Figure 3). Reducing the number of days to slaughter reduced the emissions from
368 enteric fermentation and manure. This highlights the importance of high production efficiency
369 (i.e., growth rates) in beef production, as found in other studies (e.g., Capper, 2011;
370 Wiedemann et al., 2015).

371 A reduction of Ym with increased grass silage quality have been found in feeding experiments
372 in cattle, similar to what is assumed in HolosNor. Warner et al. (2016) investigated grass silages
373 cut at three stages of maturity (early, mid and late maturity) at two levels of nitrogen
374 fertilization (65 vs. 150 kg N/ha). They found that maturity stage influenced Ym. For example
375 Ym values of 6.8, 7.2 and 7.1% were found for the early, mid and late maturity grass silages at
376 the low level of nitrogen fertilization, respectively. The diet consisted of 20% compound feed.
377 Similarly, Brask et al. (2013) compared two grass silages (early first cut vs. late first cut)
378 supplemented with two levels of fat in the concentrates. Ym for the early and late cut fed with
379 the control (ie., low-fat) concentrate were 6.4 and 6.9%.

380

381 4.3 Total emissions

382 Even though the effect of grass silage quality on emission intensity is clear, the effects of
383 improved grass silage quality on the total emissions from the cattle population is not as obvious
384 due to the relationship between milk yield per cow, beef production from the dairy enterprise
385 and the need for suckler cows. The lowest emission intensities for milk and beef were found
386 for H1C8 and H1 with concentrates, respectively (Figure 1 and Figure 3). Interestingly, this
387 combination did not result in the lowest total greenhouse gas emission (Figure 5), which was
388 found for H1C0 combined with concentrates for bulls (only minor differences with no
389 concentrates for bulls). This was because of the higher need for suckler cows to meet the
390 production target of beef for H1C8 (Figure 4). H1C0 for dairy cows combined with
391 concentrates for bulls, gave the lowest total greenhouse gas emissions due to a combination of
392 low milk yield per cow (Table 1) giving high beef production from the dairy enterprise due to
393 a high need for dairy cows (Figure 4), a low emission intensity for finished young bull carcass
394 (Figure 3) and a low need for suckler cows (Figure 4). The highest total emission from beef
395 and milk production was found for H3C4, without concentrates for bulls. Interestingly, milk
396 yield per cow was only slightly higher for H3C4 compared to H1C0 (5,100 vs. 5,500 kg). This
397 was mainly due to larger emissions from the dairy beef production, due to the higher emission
398 intensity of finished young bull carcass (Figure 3). In addition, there was need for more suckler
399 cows (Figure 4), and a higher emission intensity for milk (Figure 1). These results demonstrates
400 that the lowest emission intensities does not necessarily results in the lowest total emission and
401 highlights the importance of looking at both milk- and beef production in relation to each other
402 when investigating potential mitigation options, as pointed out by Åby et al. (2016). As no
403 effect of improved grass silage quality was included for the suckler beef, the total effect of high
404 grass silage quality on greenhouse gas emissions may be underestimated.

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408 **5 CONCLUSIONS**

409 Emission intensities for milk and beef were lowest for the superior H1 grass silage and highest
410 for the normal quality H3 grass silage, independent of concentrate levels. Higher concentrate
411 levels did not prevent increased emission intensities for lower grass silage quality (H3).
412 Realistic improvements in grass silage quality from H3 to H2 was shown to maintain milk
413 yields per cow at lower concentrate levels while reducing emissions intensity for milk by
414 approximately 10%. For young bull carcasses, the potential was a reduction of emission
415 intensity by 17%. Cutting the grass at an earlier maturity stage will improve grass silage quality,
416 have beneficial effects on emission intensities for milk and beef, and simultaneously reduce
417 the need for concentrates. The silage quality-concentrate combination that yielded the lowest
418 emission intensity for milk and highest milk yield (H8C0) did not result in the lowest total
419 greenhouse gas emission from the national cattle population. The link between milk yield and
420 beef production must be considered when investigating potential mitigation options for cattle.

421

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548 **TABLE 1** Effects of grass silage quality and concentrate level on milk yield per cow and milk
 549 composition

Silage quality ¹	Concentrate level ² , kg/d	305-day yield ³ , kg	Milk composition ⁴	
			fat%	protein%
<u>H1</u>	0	5100	4.14	3.15
	4	6900	4.13	3.22
	8	8100	4.09	3.28
	12	7200	3.97	3.32
<u>H2</u>	4	6275	4.26	3.20
	8	7000	4.12	3.18
	12	6950	3.96	3.36
<u>H3</u>	4	5550	3.88	3.22
	8	6300	3.95	3.28
	12	7200	3.89	3.22
	16	6775	3.95	3.23

550 ¹Harvesting time for grass silage: H1=very early, H2=early H3=normal (Randby et al., 2010,
 551 2012)

552 ²Concentrate level used in experiment (Randby et al., 2012)

553 ³Milk yield used as input in HolosNor (Bonesmo et al., 2013). Average daily milk yield in
 554 week 1-16 from Randby et al. (2012) converted to 305-day milk yield using a standard lactation
 555 curve of the Norwegian dairy cooperative TINE SA

556 ⁴Milk composition from Randby et al. (2012)

557

558 **TABLE 2** Effects of grass silage quality and concentrate supplementation on concentrate
 559 consumption from 7 months age until slaughter, slaughter age and slaughter weight of growing/
 560 fattening bulls (Randby et al., 2010)

Silage quality¹	Concentrate supplementation	Concentrate consumption, kg	Slaughter age, days	Slaughter weight, kg
H1	Without	0	450	572
	With	495	427	572
H2	Without	0	466	568
	With	498	432	577
H3	Without	0	543	572
	With	564	454	573

561 ¹See Table 1

562

563 **TABLE 3** Inputs used in HolosNor for included silage qualities and concentrate levels

Silage quality ¹	Dairy cow concentrate level ² , kg/d	Number of			Annual feed use		
		Dairy cows ³	Bulls ⁴	Ley area ⁵ , ha	Concentrate dairy cows ⁶ , FU ⁷	Concentrate bulls ⁸ , FU	Silage, kg DM
H1	0	36	21	29	0	3836	259976
	4	27	15	21	18148	9509	185898
	8	23	13	18	35011	8100	157359
	12	26	15	17	42735	9113	154005
H2	4	29	17	20	34113	3019	210405
	8	26	15	17	39999	9298	177169
	12	27	15	15	51846	9365	163180
H3	4	33	19	25	28752	2936	275367
	8	29	17	19	45059	10717	209098
	12	26	15	15	60743	9378	167566
	16	27	16	14	80964	9966	152328

564 ¹See Table 1

565 ²See Table 1

566 ³ The number of dairy cows needed to fulfil a target for annual fat-and protein-corrected milk
 567 production on farm (in total approximately 185 000 kg milk per year, equal to 26 cows with
 568 average milk yield 7100 ECM kg) based on the 305-days yields given in Table 1

569 ⁴ The number of finished young bulls was 0.57 per cow and year based on a dairy cow
 570 replacement rate of 0.3 calculated as a function of average values for slaughter age of cow, age
 571 at first calving, calving interval and calf losses (TINE, 2013)

572 ⁵ Estimated grass ley area for each treatment needed to cover the total silage requirement for
 573 dairy cows, replacement heifers and finished bulls based on estimated annual grass yield per
 574 ha of the three silage qualities

575 ⁶ Based on total energy requirements calculated as a function of milk yield (Volden, 2013), and
 576 corrected for observed silage intake (Randby et al., 2012)

577 ⁷1FU, feed unit=6,9 MJ net energy lactation

578 ⁸ Based on consumption per bull (with concentrates) in Table 2 plus 228 kg DM per bull before
 579 7 months of age (Berg & Matre, 2007)

580

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582

583

584 **Figure legends:**

585 **FIGURE 1** Emission intensity in kg CO₂-equivalents per kg fat- and protein-corrected milk
586 yield for the included silage qualities (H1, H2 and H3) and concentrate levels (0, 4, 8, 12 and
587 16 kg/d)

588 **FIGURE 2** The lowest (H1C8) and highest emission intensity (H3C4), distributed on emission
589 sources

590 **FIGURE 3** Emission intensities in kg CO₂-equivalents per kg beef carcass for the included
591 silage qualities, offered as sole feed (without concentrates) or supplemented with concentrates
592 (with)

593 **FIGURE 4** Number of dairy and suckler cows needed to meet the domestic production level
594 of milk (1500 million liters) and beef (80,000 tons) as a function of milk yield per dairy cow
595 on silage quality H1 with 0 and 8 kg/d concentrate, and H3 with 4 kg/d concentrate

596 **FIGURE 5** Total annual greenhouse gas emissions in CO₂-equivalents from milk and beef
597 from the dairy and suckler populations, for the dairy cow diets including silage quality H1 with
598 0 and 8 kg/d concentrate, and H3 with 4 kg/d concentrate. Each dairy cow diet is combined
599 with dairy bull diets of the same silage quality without or with concentrates. Emissions are
600 distributed on animal products: Milk from dairy cows, beef from dairy population (dairy cows
601 and bulls) and beef from suckler population. All six feeding regimes fulfill the domestic
602 production level of milk (1500 million liters) and beef (80,000 tons).

603