

RESEARCH ARTICLE

Can only poorer European countries afford large carnivores?

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Abstract

Background

One of the classic approaches in environmental economics is the environmental Kuznets curve, which predicts that when a national economy grows from low to medium levels, threats to biodiversity conservation increase, but they decrease when the economy moves from medium to high. We evaluated this approach by examining how population densities of the brown bear (*Ursus arctos*), gray wolf (*Canis lupus*), and Eurasian lynx (*Lynx lynx*) were related to the national economy in 24 European countries.

Methodology/Principal findings

We used forest proportions, the existence of a compensation system, and country group (former socialist countries, Nordic countries, other countries) as covariates in a linear model with the first- and the second-order polynomial terms of per capita gross domestic product (GDP). Country group was treated as a random factor, but remained insignificant and was ignored. All models concerning brown bear and wolf provided evidence that population densities decreased with increasing GDP, but densities of lynx were virtually independent of GDP. Models for the wolf explained >80% of the variation in densities, without a difference between the models with all independent variables and the model with only GDP. For the bear, the model with GDP alone accounted for 10%, and all three variables 33%, of the variation in densities.

Conclusions

Wolves exhibit a higher capacity for dispersal and reproduction than bear or lynx, but still exists at the lowest densities in wealthy European countries. We are aware that several other factors, not available for our models, influenced large carnivore densities. Based on the pronounced differences among large carnivore species in their countrywide relationships between densities and GDP, and a strikingly high relationship for the gray wolf, we suggest that our results reflected differences in political history and public acceptance of these species among countries. The compensation paid for the damages caused by the

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Data Availability Statement: Population data and notes concerning compensation for damages were collected from <http://ec.europa.eu/environment/nature/conservation/species/carnivores/pdf/>, land area and forest area at <https://data.worldbank.org/indicator/AG.LND.TOTL.K2>, and gross domestic product at <https://data.worldbank.org/indicator/NY.GDP.MKTP.CD>. Data are available as a supplementary file Data_PONE-D-17-36810R1.xlsx.

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carnivores is not a key to higher carnivore densities, but might be necessary for the presence of large carnivores to be accepted in countries with high GDP.

Introduction

Many animal populations are facing extinction risk caused by human impacts. Traditionally, economic growth has been regarded to threaten biodiversity conservation [1]. A major challenge to biodiversity conservation is to facilitate the protection of species that are valued at a global scale, but have negative value at a local scale, e.g., due to public fears and damage to livestock [2]. Large carnivores often cause economic damage and, conversely, the intensive land use often associated with economic growth is deleterious for large carnivore populations [3–5]. Globally the number of threatened mammal species by country is correlated with high socioeconomic factors [6]. On the other hand, biodiversity loss and poverty can be interrelated, whereby conservation might help tackle poverty through, for example, generating jobs in wildlife tourism [7].

One of the classic approaches in environmental economy is the environmental Kuznets curve, derived from Kuznets [8]. The curve predicts that when a national economy grows from low to medium levels, threats to biodiversity conservation increase, but they decrease when the economy moves from medium to high [1, 9–11]. There are several natural and human-mediated factors influencing large carnivore populations and population densities. One human-mediated reason explaining variations in population densities might be public economy, when it is associated with individual humans' possibilities to minimize livestock depredation and risks to human safety.

In many European countries, both the economy and large carnivore populations are stronger than they were some decades ago [12]. Here we examined how current population densities of the three most widely distributed European large carnivores, the brown bear (*Ursus arctos*), gray wolf (*Canis lupus*), and Eurasian lynx (*Lynx lynx*), are related to national economies in Europe. If the environmental Kuznets curve applies to these species, their population densities would be lowest at the mid-range of gross domestic product (GDP).

Material and methods

Data

We obtained mean wolf, bear, and lynx population abundance estimates for 2008–2011 from 24 European countries, based on expert estimates [13]. The availability of suitable habitat, a key factor for successful conservation of large carnivore populations, varies widely among countries. However, only a portion of the land area is suitable habitat. Therefore, we assumed that forest area was representative of the area of suitable habitat and per capita gross domestic product (GDP) was representative of the national economy. In this study, we treated population size [13] per land area of the country (i.e. population density as animals/1 000 km²) as a dependent variable. We also used this data compiled by the European Commission [13] to obtain the existence of systems for compensation of depredation caused by large carnivores. We obtained data on land area, the forest area, and GDP in 2011 for each country from World Bank statistics (<https://data.worldbank.org/>). Due to cultural similarities and differences between countries, some pseudoreplication could be present without taking a random factor of country group into statistical models. We divided the countries into three groups: 1) Nordic

countries, 2) eastern European countries, and 3) the remaining European countries (Western Europe, Nordic countries excluded).

Statistical analysis

By treating population density as a dependent variable, we fitted a linear model with the first- and the second-order polynomial terms of GDP (gross domestic product) for each species separately (brown bear, wolf and lynx). Then we did this with two independent variables, GDP and the proportion forest. The country group was entered as a random factor and the existence of a governmental compensation system as a fixed factor to all models. The country group or the existence of a compensation system did not affect densities significantly ($p > 0.10$) in any model, and therefore, they were excluded from the final models.

The models were built after log-transformation of population density, because the log-transformed distributions fit better, and the transformed values were closer to the normal distribution. We presented parameter estimates and tests in the log-transformed scale, but the predicted values after transforming them back to the original scale. The biases caused by the transformation to the original scale were corrected using an empirical bias correction presented by Snowdon [14]. The correction was based on the ratio of the observed mean of the response and the mean of the exponent of the predicted values.

The basic form of the linear models could be described as:

$$\text{Log}(\text{population density}) = \text{constant} + \beta_1 \text{GDP} + \beta_2 \text{GDP}^2 + \beta_3 \text{forest proportion} + \varepsilon,$$

where β_1 , β_2 and β_3 are the fixed coefficients, GDP denotes the gross domestic product per capita, forest proportion denotes the proportion of forest area of the total land area, and ε is residual variation. If the second-order term of the GDP influenced the parabola shape of the predicted curve in a well-fitted model with the GDP as the predictor, the Kuznets curve would be supported. For the lynx, we built post hoc models also without GDP² to evaluate the effect of GDP², because the scatterplot with the second-order term (Fig 1F) indicated an increase in population density from medium to high GDP.

The models were computed using R function `lm` [15]. In premodeling, the random factor of country group (western and eastern Europa, Nordic countries) and the existence of a compensation system were included in the models by using R package `lme4`, and its function `lmer` [16]. The variances of the country group were tested using R package `lmerTest` and its function `rand` [17].

Results

Increasing GDP was related to lower densities of brown bear and wolf (Fig 1).

Models both with and without forest proportion supported this observation (Fig 1). However, the variation in lynx density seemed to be virtually independent of GDP (Fig 1).

Based on the adjusted r-squares, the full model accounted for 81.8% and the model without forest proportion 82.3% for wolf densities, 32.9 and 10.4% for bear densities and 20.3 and 4.2% for lynx densities, respectively (Table 1).

Models without the second-order term (GDP²) accounted for only slightly less of the variation in lynx density ($\Delta\text{AIC} = 2.78$) [18] than models with GDP², due to a small increase in densities from medium to high GDP (Fig 1, Table 1).

Several factors may contribute large carnivore densities, but it is noteworthy that GDP alone accounted for about 80% of the variation in wolf population densities and did not leave explanatory room for the other independent variable, proportion of forest, in the model (Table 1). For the brown bear, on the other hand, proportion of forest had a greater

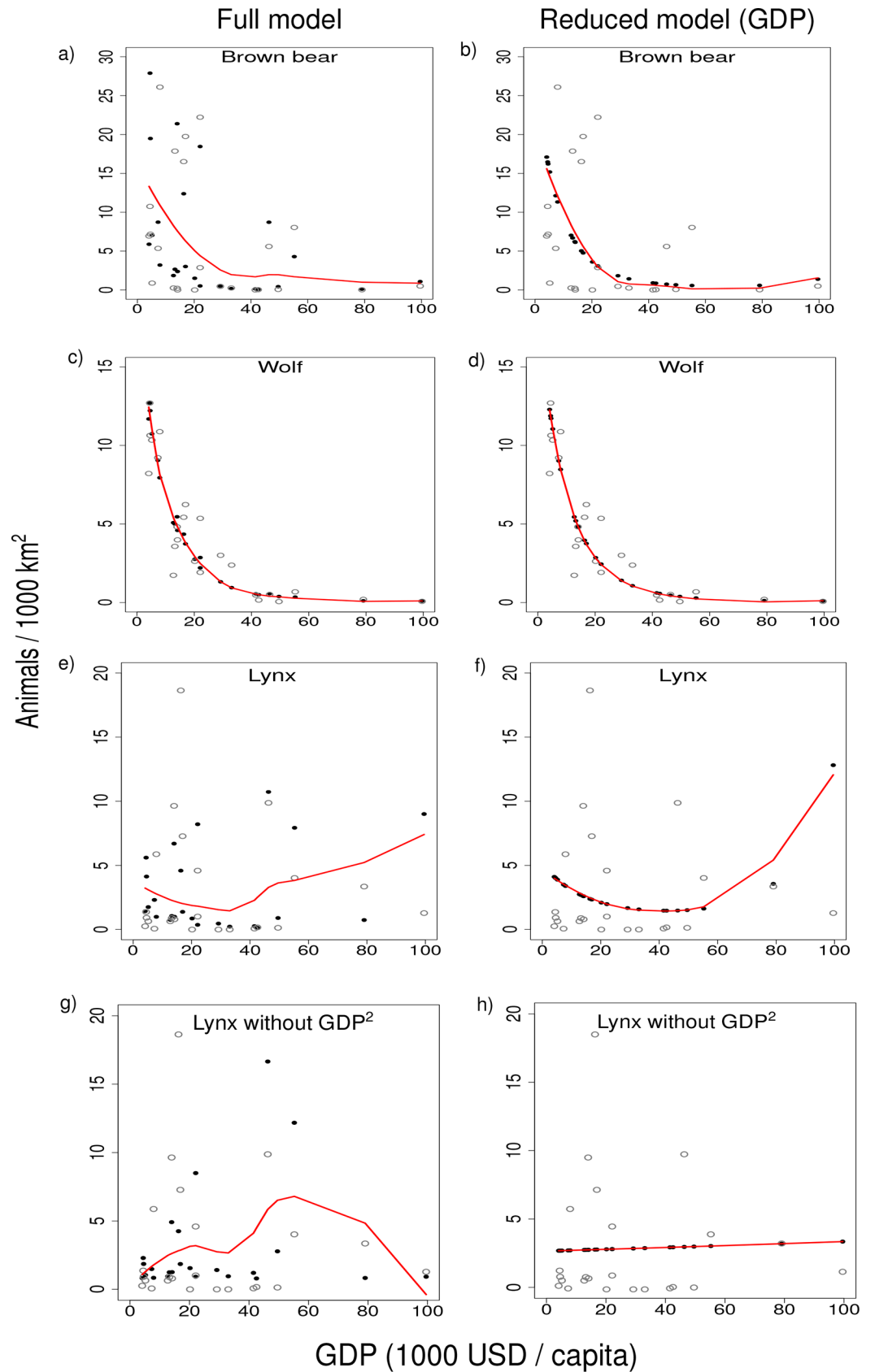


Fig 1. Relationships of per capita gross domestic product (GDP) and population densities of three large carnivore species in Europe. The full model include both forest proportion and GDP as independent variables, the reduced model only GDP. For the lynx, models without the second-order term (GDP^2) are also shown. Predicted values are shown as black circles and observed ones as open circles. The red smoothed curves are based on the predicted values.

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contribution to explaining densities than GDP (Table 1). The proportion of forest and GDP accounted for 30%, but GDP alone for only 10% of the variation in bear densities (Table 1). Although the proportion of forest was significant in the first model for lynx (Table 1), the adjusted r-squares in the lynx models suggested that the independent variables, GDP and forest proportion, had no clear contribution to the country-specific lynx densities in Europe. For the brown bear model, predictions for the forest proportions of 0.3, 0.5, and 0.7 were population densities of 0.28, 2.58 and 23.92 brown bears 1000 km⁻², respectively. The predictions for lynx based on the first model (with the second-order term of GDP) were 0.27, 1.87, and 12.77 lynx 1000 km⁻² with forest proportions of 0.3, 0.5, and 0.7, respectively.

Discussion

Our analyses demonstrated that large carnivore densities in Europe generally are lowest in the wealthy countries. However, the relationship between GDP and population density clearly varied by carnivore species. The relationship was very strong for the gray wolf, significantly weaker for the brown bear, and did not exist for the Eurasian lynx. These patterns are likely explained by the public image of different carnivore species and political history in Europe.

We did not find any indication of an environmental Kuznets U-shape curve between large carnivore density and national economy in Europe. On the contrary, we found a strong

Table 1. The nonadjusted and adjusted r-squares, parameter estimates, and tests for the two linear models (full models and reduced models without proportion of forest) explaining the population density per 1000 km² of brown bears, gray wolves, and Eurasian lynx in European countries. Two alternative models are presented for the lynx (Lynx 1 and Lynx 2), to define the effect of the second-order term of GDP (with and without GDP^2).

Variable	Full model			Reduced model		
	Estimate (SE)	p-value	R ²	Estimate (SE)	p-value	R ²
Brown bear			0.416/0.329			0.182/0.104
Intercept	-1.46 (1.58)	0.364		2.03 (1.14)	0.090	
GDP	-0.12 (0.07)	0.085		-0.19 (0.07)	0.102	
GDP ²	1.59e-3 (0.68e-3)	0.030		0.87e-3 (0.73e-3)	0.240	
Forest prop.	11.12 (3.93)	0.010				
Wolf			0.842/0.818			0.838/0.823
Intercept	2.62 (0.49)	<0.001		1.99 (0.33)	<0.001	
GDP	-0.11 (0.20)	<0.001		-0.09 (0.20)	<0.001	
GDP ²	0.56e-3 (0.21e-3)	0.014		0.40e-3 (0.20e-3)	0.075	
Forest prop.	0.82 (1.21)	0.507				
Lynx 1			0.307/0.203			0.049/0.042
Intercept	-2.71 (1.42)	0.070		1.30 (0.88)		
GDP	-0.12 (0.06)	0.054		-1.91 (0.07)		
GDP ²	0.12e-3 (0.61e-3)	0.049		0.11e-3 (0.56e-3)		
Forest prop.	9.61 (3.52)	0.013				
Lynx 2			0.155/0.074			0.000/0.000
Intercept	-3.10 (1.51)	0.053		0.70 (0.65)	0.918	
GDP	-0.00 (0.02)	0.892		-0.00 (0.18)	0.891	
Forest prop.	6.91 (3.53)	0.064				

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negative correlation between GDP and wolf density, a moderate negative correlation between GDP and bear density, and no correlation between GDP and lynx density. If the recent expansion and increase of wolf populations in Western Europe continues [12] or the rank between national economies changes, this pattern may also change in the coming decades. Brown bear and lynx densities will change more slowly, especially bear densities, owing to the species' low rates of reproduction and dispersal [19].

Wolves cause the highest per capita losses among carnivores to the sheep and domestic reindeer (*Rangifer tarandus*) industries on the European scale [13,20]. Bears and wolves may attack humans too, but attacks are exceptionally rare [21–24]. The fear of wolves is common due both to real and perceived threats [22,23].

In Europe, wolf population numbers are generally largest in the eastern parts of the continent [12,25]. One potential reason for this is associated with political history. Motivational aspects of wildlife management might be linked to property rights and landownership [26]. For example, the rights to hunt ungulates, the primary prey of wolves, are bound to landownership in Scandinavia, Finland, and Germany [27–30] but not in many former socialist countries, such as the Baltic countries and Poland [31,32]. Mean wolf densities in the Baltics and Poland are considerably higher than in Scandinavia, Finland, and Germany, where ungulate hunters may potentially be motivated to control the wolves as competitors for ungulates. This might be especially true for Scandinavia and Finland, where mean ungulate biomass is low and where producing food is still a major focus of ungulate management [33]. In Scandinavia, acceptance of the illegal hunting of large carnivores is high in areas with strong hunting traditions [34]. Furthermore, citizens of some former socialist countries might have had fewer means to control large carnivore populations than in western democracies, due to limited access to effective firearms.

Another potential reason for the present distribution pattern of large carnivores regards the geography or source populations, especially those within the former Soviet Union, which provided continuous dispersal to other Eastern European countries. Political changes in former socialist countries may also have affected large carnivore populations. In Poland, lynx mortality due to poaching was higher during the low GDP years following the collapse of communism and lower in subsequent years, when GDP had risen to a higher level [35]. In addition, different species may react differently to economic and social perturbations associated with the drastic changes in Eastern Europe. In Russia, for example, numbers of brown bear and lynx decreased whereas wolf numbers increased, following the collapse of the Soviet Union [36,37].

All large carnivores are mobile and especially the gray wolf has a high dispersal potential and colonizes new areas relatively easily. Recently, wolves have naturally reestablished breeding populations in Finland, Sweden, Norway, France, and Germany [12, 37, 38, 39]. They outnumber brown bears in most Central European and Mediterranean countries, but occur at lower numbers than bears in Estonia, Scandinavia, and Finland (12). Suitable habitat in some Western European countries, such as the United Kingdom, Netherlands, Denmark, and Belgium, is too small and isolated for the establishment of a wolf population and the species was exterminated there several hundred years ago [25]. However, wolves were also exterminated from Scandinavia [38], where human densities are among the lowest in Europe.

Compensation paid for the damages are lowest in poorer countries, even when corrected for per capita GDP (Kojola et al., unpublished data). Compensation for the damages does not unequivocally enhance the conservation of large carnivore populations [40,41], but may be a requirement for large carnivore conservation in wealthier countries in Europe, into which large carnivore populations have recently expanded [12] and where people are therefore less used to their presence. Paying for their presence may promote conservation of large carnivore

populations [42, 43], although there is also evidence that predator poaching is influenced more strongly by social than economic factors [44].

Our results reflected differences in political history and public acceptance of three large carnivore species in Europe. Although large carnivores, including the gray wolf, have been increasing recently in many wealthy countries [12], a strikingly high negative relationship still prevails between GDP and wolf densities. Relationships to GDP may weaken in the future if conservation policies will favor higher large carnivore densities in wealthy countries. The ongoing population growth of large carnivores in countries with high GDP will provide a test of how well wealthy countries will manage the conservation of large carnivores. The compensation paid for the damages caused by the carnivores is not a key to allowing higher carnivore densities, but might be necessary for the presence of large carnivores to be accepted in countries with a high national economy.

Supporting information

S1 Table. Data used in this study.
(XLSX)

Author Contributions

Conceptualization: Ilpo Kojola, Timo Helle.

Data curation: Ville Hallikainen.

Investigation: Ilpo Kojola.

Writing – original draft: Ilpo Kojola.

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References

1. Czech B (2008) Prospects for reconciling the conflict between economic growth and biodiversity conservation with technological progress. *Cons Biol* 22: 1389–1398.
2. Dickman J, Macdonald EA, Macdonald DW (2011) A review of financial instruments to pay or predator conservation and encourage human-carnivore coexistence. *PNAS* 108: 13937–13944. <https://doi.org/10.1073/pnas.1012972108> PMID: 21873181
3. Woodroffe R (2000) Predators and people: using human densities to interpret declines of large carnivores. *Anim Cons* 3: 165–173.
4. Treves A, Karanth UK (2003) Human-carnivore conflict and perspectives on carnivore management worldwide. *Cons Biol* 17: 1491–1499.
5. Cardillo M, Purvis C, Sechrest W, Gittlemann JL, Bielby J, Mace GM (2004) Human population density and extinction risk in the world's carnivores. *PLOS Biol* 2: e197. <https://doi.org/10.1371/journal.pbio.0020197> PMID: 15252445
6. Polaina E., Gonzalez-Suarez M., and Revilla E. 2015. Socioeconomic correlates of global mammalian conservation status. *Ecosphere* 6:146.
7. Adams W, Aveling R, Brockington DB, Dickson J, Elliott J, Hutto D et al. (2004) Biodiversity conservation and the eradication of poverty. *Science* 306: 1146–1149. <https://doi.org/10.1126/science.1097920> PMID: 15539593
8. Kuznets S (1955) Economic Growth and Income Inequality. *Am Econ Rev* 45: 1–28.
9. Selden TM, Song D (1994) Environmental Quality and Development: Is there a Kuznets curve for air pollution emissions? *J Env Econ Manage* 27: 147–162.
10. Koop G, Tole L (1999) Is there an environmental Kuznets curve for deforestation? *J Dev Econ* 58: 231–244.
11. Stern DI (2004) The Rise and fall of the environmental Kuznets curve. *World Dev* 32: 1419–1432.

12. Chapron G, Kaczensky P, Linnell JDC, von Arx M, Huber D, Andren H et al. (2014) Recovery of large carnivores in Europe's modern human-dominated landscapes. *Science* 346: 1517–1519. <https://doi.org/10.1126/science.1257553> PMID: 25525247
13. Kaczensky P, Chapron G, von Arx M, Huber D, Andrén H, Linnell J (2013) Status, Management and Distribution of Large Carnivores, Bear, Lynx, Wolf and wolverine in Europe. (http://ec.europa.eu/environment/nature/conservation/species/carnivores/pdf/task_1_part2_species_country_reports.pdf 2013).
14. Snowdon P (1991) A ratio estimator for bias correction in logarithmic regressions. *Can J For Res* 21: 720–724.
15. R Core Team (2016) R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. URL <https://www.R-project.org/>.
16. Bates D, Maechler M, Bolker B, Walker S (2016) Fitting linear mixed-effects models using lme4. *J Stat Softw* 67: 1–48.
17. Kuznetsova A, Brockhoff PB, Christensen RHB (2016) lmerTest: tests in linear mixed effects models. R package version 2.0–30. <https://CRAN.R-project.org/package=lmerTest>.
18. Burnham KP, Anderson DR (2004) Multi model interference: understanding AIC and BIC in model selection. *Sociol Met Res* 33: 261–304.
19. Swenson JE, Sandegren F, Söderberg A (1998) Geographic expansion of an increasing brown bear population: evidence for presaturation dispersal. *J Anim Ecol* 67: 819–826.
20. Nieminen M, Norberg H, Majjala V (2013) Calf mortality of semi-domesticated reindeer (*Rangifer tarandus tarandus*) in the Finnish reindeer-herding area. *Rangifer, Spec Issue* 21: 79–89.
21. Løe J, Røskoft E (2004) Large carnivores and human safety: a review. *Ambio* 33: 283–288. PMID: 15387060
22. Linnell JDC, Solberg E, Brainerd S, Liberg O, Sand H, Wabakken P et al. (2003) Is the fear of wolves justified? A Fennoscandian perspective. *Acta Zool Lituonica* 13: 34–40.
23. Ginsberg J (2001) Setting priorities for carnivore conservation: what makes carnivores different? In: Gittleman JL, Funk SM, Macdonald D, Wayne RK, editors. *Carnivore conservation*. Cambridge University Press, Cambridge, UK. pp. 498–523.
24. Penteriani V, Delgado M, Pinchera F, Naves J, Fernandez-Gil J, Kojola I et al. (2016) Human behaviour can trigger large carnivore attacks in developed countries. *Sci Rep* 6: 20552. <https://doi.org/10.1038/srep20552> PMID: 26838467
25. Boitani L (2003) Wolf conservation and recovery. In: Mech LD, Boitani L, editors. *Wolves. Behavior, ecology and conservation*. The University Chicago Press. Chicago, London. pp. 317–340.
26. Naughton-Treves L (1999) Whose animals? A history of property rights to wildlife in Toro, western Uganda. *Land Degrad Dev* 10: 311–328.
27. Andersen R, Lund E, Solberg EJ, Saether B-E (2010) Ungulates and their management in Norway. In: Apollonio M, Andersen R, Putman R, editors. *European ungulates and their management in the 21st century*. Cambridge University Press, Cambridge, UK. pp. 14–36.
28. Liberg O, Bergström R, Kindberg J, von Essen H (2010) Ungulates and their management in Sweden. In: Apollonio M, Andersen R, Putman R, editors. *European ungulates and their management in the 21st century*. Cambridge University Press, Cambridge, UK. pp 37–70.
29. Ruusila V, Kojola I (2010) Ungulates and their management in Finland. In: Apollonio M, Andersen R, Putman R, editors. *European ungulates and their management in the 21st century*. Cambridge University Press, Cambridge, UK. pp. 86–102.
30. Wotchikowsky U (2010) Ungulates and their management in Germany. In: Apollonio M, Andersen R, Putman R, editors. *European ungulates and their management in the 21st century*. Cambridge University Press, Cambridge, UK. pp. 201–222.
31. Andersone-Tilley Z, Balčiauskas L, Ozolins J, Randveer T, Tonisson J (2010) Ungulates and their management in the Baltics (Estonia, Latvia, Liettua). In: Apollonio M, Andersen R, Putman R, editors. *European ungulates and their management in the 21st century*. Cambridge University Press, Cambridge, UK. pp. 103–128.
32. Wawrzyniak P, Jedrzejewski W, Jedrzejewska B, Borowik T (2010) Ungulates and their management in Poland. In: Apollonio M, Andersen R, Putman R, editors. *European ungulates and their management in the 21st century*. Cambridge University Press, Cambridge, UK. pp. 223–242.
33. Putman R, Andersen R, Apollonio M (2011) Introduction. In: Putman R, Apollonio M, Andersen R, editors. *Ungulate management in Europe. Problems and practices*. Cambridge University Press Cambridge, UK. pp.1–11.

34. Gangaas KE, Kaltenborn BP, Andreassen HP (2013) Geo-spatial aspects of acceptance of illegal hunting of large carnivores in Scandinavia. *PLoS ONE* 8: e68849. <https://doi.org/10.1371/journal.pone.0068849> PMID: 23894353
35. Kowalczyk R, Gorny M, Schmidt K (2015) Edge effect and influence of economic growth on Eurasian lynx mortality in the Bialowieza Primeval Forest, Poland. *Mammal Res* 60: 3–8.
36. Bragina EV, Ives AR, Pidgeon AM, Baskin LM, Gubar YB, Piquer-Rodrigues M et al. (2015) Rapid declines of large mammal populations after the collapse of the Soviet Union. *Cons Biol* 29:844–853.
37. Bragina EV, Ives AR, Pidgeon AM, Balčiauskas L, Csányi S, Khoyetsky P et al. (2018) Wildlife population changes across Eastern Europe after the collapse of socialism. *Frontiers Ecol Env*. <https://doi.org/10.002/>
38. Wabakken PH, Sand H, Bjärvall A (2001) The recovery, distribution, and population dynamics of wolves on the Scandinavian peninsula, 1978–1998. *Can J Zool* 79: 710–725.
39. Kojola I, Helle P, Heikkinen S, Linden H, Paasivaara A, Wikman M (2014) Tracks in snow and population size estimation: wolf in Finland. *Wildl Biol* 20: 279–284.
40. Naughton-Treves L, Grossberg R, Treves A (2003) Paying for tolerance: Rural citizens' attitudes toward wolf depredation and compensation. *Cons Biol* 17: 1500–1511.
41. Bulte EH, Rondeau D (2006) Why compensating wildlife damages may be bad for conservation. *J Wildl Manage* 69: 14–19.
42. Nelson F (2009) Developing payments for ecosystem services approaches to carnivore conservation. *Hum Dim Wildl* 14: 381–392
43. Persson J., Rauset GR, Chapron G (2015) Paying for an endangered predator leads to population recovery. *Cons Lett* 8: 345–350.
44. Treves A, Bruskotter J (2014) Tolerance for predatory wildlife. *Science* 344: 476–477. <https://doi.org/10.1126/science.1252690> PMID: 24786065